

# Unit-level and area-level small area estimation under heteroscedasticity using digital aerial photogrammetry data

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## Abstract

1 In many applications, estimates are required for small sub-populations with so few (or no) sample  
2 plots that direct estimators that do not utilize auxiliary variables (e.g. remotely sensed data) are  
3 not applicable or result in low precision. This problem is overcome in small area estimation (SAE)  
4 by linking the variable of interest to auxiliary variables using a model. Two types of models can  
5 be distinguished based on the scale on which they operate: i) Unit-level models are applied in the  
6 well-known area-based approach (ABA) and are commonly used in forest inventories supported  
7 by fine-resolution 3D remote sensing data such as airborne laser scanning (ALS) or digital aerial  
8 photogrammetry (AP); ii) Area-level models, where the response is a direct estimate based on  
9 a sample within the domain and the explanatory variables are aggregated auxiliary variables,  
10 are less frequently applied. Estimators associated with these two model types can make use of  
11 sample plots within domains if available and reduce to so-called synthetic estimators in domains  
12 where no sample plots are available. We used both model types and their associated model-based  
13 estimators in the same study area with AP data as auxiliary variables. Heteroscedasticity, i.e.  
14 for continuous dependent variables typically an increasing dispersion of residuals with increasing  
15 predictions, is often observed in models linking field- and remotely sensed data. This violates the  
16 model assumption that the distribution of the residual errors is constant. Complying with model  
17 assumptions is required for model-based methods to result in reliable estimates. Addressing  
18 heteroscedasticity in models had considerable impacts on standard errors. When complying

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19 with model assumptions, the precision of estimates based on unit-level models was, on average,  
20 considerably greater (29%-31% smaller standard errors) than those based on area-level models.  
21 Area-level models may nonetheless be attractive because they allow the use of sampling designs  
22 that do not easily link to remotely sensed data, such as variable radius plots.

*Keywords:* **Keywords:** Forest inventory, model-based inference, synthetic estimator, variance estimation, image matching

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## 23 1. Introduction

24 Fine-resolution remotely sensed data such as 3D point clouds acquired using airborne laser  
25 scanning (ALS) or digital aerial photogrammetry (AP) can be utilized for estimating forest pa-  
26 rameters with individual tree crown approaches (ITC), area-based approaches (ABA), or domain-  
27 level approaches (DLA). The three approaches can be linked to two model types (unit-level and  
28 area-level) and their associated model-based estimators in small area estimation (SAE). The aim  
29 is typically to provide estimates of forest parameters on the level of stands or other domains.  
30 For example, main outcomes of an inventory using the ABA are stand-level mean timber vol-  
31 ume estimates resulting from averaging the grid-cell predictions per stand. Today, ITC and  
32 ABA are commonly applied in operational forest management inventories (Næsset, 2014) while  
33 domain-level approaches (DLA) have mostly been used in research studies (van Aardt et al.,  
34 2006; Goerndt et al., 2011).

35 Unit-level models are applied in ITC and ABA because the variable of interest and the  
36 explanatory variables are available on the level of population units such as geo-located trees or  
37 field sample plots (Rao, 2003, Ch. 5.3). Area-level models are applied in DLA if variables of  
38 interest are estimates based on a sample within a domain (Rao, 2003, Ch. 5.2). The explanatory  
39 variables in area-level models are obtained by aggregating the auxiliary variables to the domain  
40 level, for example by calculating the mean of the ALS heights within each forest stand. The  
41 advantage of area-level models is that no exact geo-locations are required for sample plots which  
42 can be also of interest in cases where plot coordinates are confidential. Furthermore, it allows  
43 the use of plot designs that cannot be exactly matched to remotely sensed data. For example,  
44 it may be difficult to link a line transect, as used in distance sampling, to remotely sensed data  
45 and to tessellate the study area according to the transect as required for unit-level estimators  
46 (Bäuerle et al., 2009). Similar issues arise with variable radius plots although workarounds may  
47 be feasible (Kirchhoefer et al., 2017). In that way, area-level models may simplify the use of

48 remote sensing techniques, such as mobile laser scanning (Saarela et al., 2017) or UAV-based  
49 data acquisitions (Nevalainen et al., 2017), as reference data in larger scale applications.

50 Inference in remote sensing-supported forest inventories, i.e. the estimation of a population  
51 parameter and the associated uncertainty, using unit-level estimators has recently received more  
52 attention (Mandallaz, 2013; McRoberts et al., 2014; Saarela et al., 2015, 2016; Gregoire et al.,  
53 2016; Chen et al., 2016; Mauro et al., 2016). Also, area-level estimators are becoming more  
54 studied (Goerndt et al., 2011, 2013; Boubeta et al., 2015; Magnussen et al., 2017). However,  
55 unit- and area-level estimators are seldom compared (Hidiroglou and You, 2016). Mauro et al.  
56 (2017b) compared unit- and area-level estimators in an ALS-supported forest inventory where  
57 the domains were forest stands aggregated to management units larger than 4 ha. They found  
58 that the root mean squared errors (RMSE) of area-level estimates were, on average, between 1.3  
59 (Lorey's height) and 2.8 (timber volume) larger than RMSE of unit-level estimates.

60 For continuous dependent variables such as timber volume, heteroscedasticity typically man-  
61 ifests itself as an increasing dispersion of residuals with increasing predictions. It is frequently  
62 observed in linking models for remote-sensing supported forest inventories (e.g., Rahlf et al.,  
63 2014; Saarela et al., 2016). Heteroscedasticity may be caused by natural phenomena but can  
64 also indicate an omitted explanatory variable or a mis-specified model shape (e.g., linear instead  
65 of curvilinear relationship). How to handle heteroscedasticity in SAE is still an active field  
66 of research and did not receive much attention in area-level estimation. In unit-level estima-  
67 tion, Militino et al. (2006) used the number of sample units within a domain for considering  
68 heteroscedasticity. This choice, however, does not allow any synthetic estimates that are often  
69 required in forest inventories. Mauro et al. (2017b) chose one of three transformations of the most  
70 important explanatory variable based on goodness of fit criteria and visual inspections of the  
71 residuals to model the heteroscedasticity. An alternative was suggested by Jiang and Nguyen  
72 (2012) who assumed that the data come from different super-populations by categorizing the  
73 sample units into few groups. The mean squared error is then a function of the empirical resid-  
74 uals within each group. This has, however, the disadvantage that the continuous nature of the  
75 heteroscedasticity is categorized and the number of super-populations has to be selected.

76 The aim of this study was to compare area-level and unit-level models and associated model-  
77 based estimators in a case study with forest inventory data using AP as auxiliary variables.  
78 Specifically, we analyze the consequences of heteroscedasticity because realistic uncertainty as-  
79 sessment of estimators associated with area-level and unit-level models depends on compliance

80 with model assumptions. We study the case where several field sample plots are available within  
81 stands which are used in linear models linking the variable of interest with auxiliary information.  
82 Synthetic estimators, i.e. aggregates of model predictions, are applied in stands without sample  
83 plots (Breidenbach et al., 2015; Magnussen et al., 2016).

## 84 2. Methods

### 85 2.1. The direct estimator

86 The aim is to estimate the population mean of some variable of interest (e.g., timber volume)  
87 for each of  $i = 1, \dots, M$  domains (small areas). Each domain is composed of  $j = 1, \dots, N_i$   
88 population units, such that

$$\bar{Y}_i = \frac{1}{N_i} \sum_{j=1}^{N_i} y_{ij} \quad (1)$$

89 where  $y_{ij}$  is the variable of interest of the  $j$ -th population unit within the  $i$ -th domain and  $N_i$  is  
90 the known number of population units within the  $i$ -th domain.

91 The estimate is denoted  $\hat{Y}_i^D$  and can be calculated from the sample units (e.g., sample plots)  
92 for the  $m \leq M$  domains (e.g., forest stands). This estimator is denoted direct (D) and is usually  
93 a Horvitz-Thompson estimate of the mean. Assuming simple random sampling (SRS) within a  
94 domain, the estimator is given by

$$\hat{Y}_i^D = \frac{1}{n_i} \sum_{j=1}^{n_i} y_{ij} \quad (2)$$

95 where  $y_{ij}$  is the  $j$ -th sample unit in domain  $i$ ,  $i = 1, \dots, m$ ,  $j = 1, \dots, n_i$ , and  $n_i$  is the number  
96 of sample units within domain  $i$ . Its variance is estimated by

$$\hat{\sigma}_i^2 = F_i \frac{s_i^2}{n_i} \quad (3)$$

97 where

$$F_i = \frac{N_i - n_i}{N_i} \quad (4)$$

98 is the finite population correction that is close to one and can thus be ignored for large populations  
99 where the sampling fraction  $f_i = \frac{n_i}{N_i}$  is small, and

$$s_i^2 = \sum_{j=1}^{n_i} (y_{ij} - \bar{y}_i)^2 / (n_i - 1) \quad (5)$$

100 is the sample variance. Domain totals and their variances can be estimated by multiplying the  
101 mean estimate with the domain size  $N_i$  and the variance estimate with  $N_i^2$ .

## 102 2.2. Notation for the use of auxiliary variables

103 We assume that for each population unit (e.g., pixels or grid cells) a vector of  $p$  explanatory  
104 variables  $\mathbf{x}_{ij}, j = 1, \dots, N_i$  (auxiliary variables), obtained from remotely sensed data, is available  
105 (wall-to-wall). Consequently, the explanatory variables are available for each sample unit  $\mathbf{x}_{ij}, i =$   
106  $1, \dots, n_i$ , and can be used in unit-level models (next section).

107 The mean of the explanatory variables over all population units within a domain ( $\mathbf{x}_{ij}, j =$   
108  $1, \dots, N_i$ ) will be denoted  $\bar{\mathbf{x}}_{iP}$  and can be used for area-level or unit-level estimates on domain  
109 level. The mean of the explanatory variables over the  $N_i - n_i$  population units not sampled  
110 within a domain ( $\mathbf{x}_{ij}, j = 1, \dots, N_i - n_i$ ) will be denoted  $\bar{\mathbf{x}}_{iR}$ . In forest inventories, it will often  
111 be the case that some domains (stands) do not contain any sample plots. Explanatory variables  
112 for those domains are still available and will be used for so-called synthetic estimates. In this  
113 case, synthetic estimates are aggregate statistics, such as mean or sum, of the model predictions  
114 over all population units in a domain.

115 Following Rao and Molina (2015), the term variance will be used for design-based estimators  
116 (i.e., the direct estimator, section 2.1), while the term mean squared error (MSE) will be used for  
117 model-based estimators which are not necessarily design-unbiased. The square root of variance or  
118 MSE estimates will be denoted standard error (SE) (Rao and Molina, 2015, p. 187). The software  
119 implementation of the methods described below is available as an R-package (Breidenbach, 2013).

## 120 2.3. Unit-level models and estimators

121 To estimate the population mean (Rao and Molina, 2015, Ch. 7), unit-level estimators use  
122 the nested-error linking model

$$123 \quad y_{ij} = \mathbf{x}_{ij}^T \boldsymbol{\beta} + v_i + \epsilon_{ij}, \quad v_i \sim N(0, \sigma_v^2), \quad \epsilon_{ij} \sim N(0, k_{ij}^2 \sigma_e^2), \quad i = 1, \dots, m, \quad j = 1, \dots, n_i \quad (6)$$

123 where  $\boldsymbol{\beta}$  is a vector of coefficients,  $v_i$  is an independently identically distributed random intercept  
124 with variance  $\sigma_v^2$  which is assumed to be independent of the residuals  $\epsilon_{ij}$  which are assumed to  
125 be mutually independent with variance  $\sigma_e^2$ , and  $k_{ij}$  is a known constant used for capturing  
126 heteroscedasticity. For  $k_{ij} = 1$ , the model is suitable for homoscedastic residual variances and all

127 further estimators simplify considerably (Prasad and Rao, 1990; Breidenbach and Astrup, 2012).  
 128 Transformed residuals and standardized residuals ( $\sigma_v$ - and  $\sigma_e$ -residuals) can be used to test the  
 129 model assumptions of homoscedasticity and normality (see AppendixA.1.1 for details).

130 For large populations and negligible sampling fractions (Battese et al., 1988), the residual  
 131 error mean assumes its expected value (zero) and the EBLUP estimator (empirical best linear  
 132 predictor) of the domain mean is

$$\hat{\mu}_i^{\text{UE}} = \bar{\mathbf{x}}_{iP}^T \hat{\boldsymbol{\beta}} + \hat{v}_i = \hat{\mu}_i^{\text{US}} + \hat{v}_i \quad (7)$$

133 where the superscript UE denotes the Unit-level EBLUP estimator and the superscript US de-  
 134 notes the Unit-level (EBLUP) Synthetic estimator. The synthetic estimator

$$\hat{\mu}_i^{\text{US}} = \bar{\mathbf{x}}_{iP}^T \hat{\boldsymbol{\beta}} \quad (8)$$

135 is the mean over all predictions within a domain which makes the estimator applicable for domains  
 136 without samples.

137 The weight  $\hat{\gamma}_i$  is the ratio of the unexplained among-domain variability (the random-effect  
 138 variance,  $\hat{\sigma}_v$ ) and the total variability

$$\hat{\gamma}_i = \frac{\hat{\sigma}_v^2}{\hat{\sigma}_v^2 + \hat{\sigma}_e^2/a_i} \quad (9)$$

139 where  $a_i = \sum_{j=1}^{n_i} a_{ij}$  and  $a_{ij} = k_{ij}^{-2}$ . Under homoscedasticity ( $k_{ij} = 1$ ),  $a_i$  reduces to  $n_i$ . The  
 140 smaller the relative unexplained among-domain variability (i.e., the more variance is explained  
 141 by the fixed part of model (6)), the smaller is  $\hat{\gamma}_i$  and the more weight is given to the synthetic  
 142 estimator (i.e., the smaller is the predicted random effect  $\hat{v}_i$ ).

143 For domains with small populations (Rao and Molina, 2015, Ch. 7.1.3), sampling fractions  
 144 ( $f_i$ ) can be non-negligible and the EBLUP estimator is given by

$$\hat{Y}_i^{\text{UE}} = \hat{\mu}_i^{\text{US}} + \omega_i (\hat{Y}_i^D - \hat{\mu}_i^{\text{US}}) \quad (10)$$

145 where  $\omega_i = f_i + (1 - f_i)\hat{\gamma}_i$  and  $\hat{Y}_i^D$  is the sample mean (eq. 2). The symbol  $\hat{Y}_i^{\text{UE}}$  is used here  
 146 rather than  $\hat{\mu}_i^{\text{UE}}$  because the sample units are considered in the estimate.

147 The uncertainty (MSE) of unit-level estimators results from estimating the random effects,

148 the fixed model parameters, and the residual error (see AppendixA.1.2 for details).

149 *2.4. Area-level models and estimators*

150 In the area-level estimators (Rao and Molina, 2015, Ch. 6), the model that links the auxiliary  
 151 variables to the direct mean estimate is described as a mixed-effects model

$$\hat{Y}_i^D = \bar{\mathbf{x}}_{iP}^T \boldsymbol{\beta} + b_i v_i + \epsilon_i, \quad v_i \sim N(0, \sigma_v^2), \quad \epsilon_i \sim N(0, \sigma_i^2), \quad i = 1, \dots, m \quad (11)$$

152 where  $v_i$  is a random intercept,  $b_i$  is a known constant for considering heteroscedasticity and  $\sigma_i^2$   
 153 is the variance of the direct estimator. With  $b_i = 1$ , the model is suited for homoscedastic data.  
 154 The direct estimator  $\hat{Y}_i^D$  is based on sample units within a domain such as eq. (2) but can be  
 155 based on any probability sampling design. This allows the use of plot designs that are difficult  
 156 to link to remotely sensed data on the population unit level. Because of the missing repetitions  
 157 of observations on domain-level, the estimation of the fixed-effects parameters  $\boldsymbol{\beta}$  is non-standard  
 158 (see AppendixA.2.1 for details on estimating fixed and random effects).

159 The area-level EBLUP estimator (superscript AE) (Fay and Herriot, 1979)

$$\tilde{\mu}_i^{\text{AE}} = \hat{\mu}_i^{\text{AS}} + b_i v_i = \bar{\mathbf{x}}_{iP}^T \hat{\boldsymbol{\beta}} + b_i v_i \quad (12)$$

160 is the weighted average of the synthetic and a direct estimate such as eq. (2).

$$\hat{\mu}_i^{\text{AE}} = \hat{\gamma}_i \hat{Y}_i^D + (1 - \hat{\gamma}_i) \hat{\mu}_i^{\text{AS}} \quad (13)$$

161 with

$$\hat{\gamma}_i = \frac{\hat{\sigma}_v^2 b_i^2}{\hat{\sigma}_v^2 b_i^2 + \hat{\sigma}_i^2}. \quad (14)$$

162 Here,  $\hat{\gamma}_i$  gives more weight to the direct estimate if its variance is relatively small and vice versa.  
 163 The more variance is explained by the fixed part of model (11), the more weight is given to the  
 164 synthetic estimator. The area-level synthetic estimator (superscript AS) can be used for domains  
 165 without observations and is given by

$$\hat{\mu}_i^{\text{AS}} = \bar{\mathbf{x}}_{iP}^T \hat{\boldsymbol{\beta}}. \quad (15)$$

166 The uncertainty (MSE) of area-level estimators results from estimating the random effect  
 167 and the fixed model parameters, and the uncertainty of the direct estimate (see AppendixA.2.2  
 168 for details).

### 169 3. Case study

#### 170 3.1. Overview

171 The study area consisted of parts of Vestfold county in south-eastern Norway with a full  
172 coverage of digital aerial photogrammetry (AP) data. Our aim was to estimate the mean timber  
173 volume scaled to per-hectare values for each of  $i = 1, \dots, m = 64$  stands (domains) from which  
174 between  $n_i = 4$  and  $n_i = 7$  population units were sampled using fixed-area sample plots. The  
175 sample plots had an area of 250 m<sup>2</sup> and trees were recorded according to the protocol of the  
176 Norwegian National Forest Inventory (NFI) for temporary plots (Landsskogtakseringen, 2008).  
177 Single tree volume was predicted from diameter at breast height and tree height recordings  
178 applying the standard models used in the NFI. Volume per hectare on plot-level was obtained  
179 by aggregating and expanding the single tree predictions. Uncertainties arising from volume  
180 models were ignored in this study. Volume per hectare,  $y_{ij}$ , ranged between 0.0 and 947.8 m<sup>3</sup>/ha  
181 on the  $n = \sum_{i=1}^{i=64} n_i = 382$  sample plots. The stands were sampled from areas with available  
182 forest management inventories (FMI) in the municipalities of Holmestrand, Lardal, and Stokke  
183 (Fig. 1).

184 A total of 30 stands was selected in Lardal with an aim to take a sample of 7 plots per stand  
185 (two stands had 6 plots). In Holmestrand and Stokke, 14 and 15 stands were selected with an  
186 aim to take a sample of 5 plots per stand (two stands had 4 plots). A constrained random sample  
187 of stand polygons that were delineated in the FMI was selected. The constraints were based on  
188 area (between 1 and 3 ha in size), and shape (avoidance of stands with a very complex outline) in  
189 order to simplify the field work. More information on the field data can be found in Breidenbach  
190 et al. (2015).

191 Potential explanatory variables were the mean and other metrics (McGaughey, 2014) of AP  
192 heights. The AP metrics were calculated for the sample plots and for square grid cells with  
193 16 m side length tessellating the study area. The grid cell size was selected to correspond  
194 approximately with the sample plot size. Image matching was performed using Socet Set 5.5.0  
195 on digital aerial images with a pixel size of 20 cm that were acquired using a Vexcel UltraCamX  
196 sensor. More information on the remotely sensed data can be found in Breidenbach and Astrup  
197 (2012).

198 The number of grid cells within a stand  $N_i$  ranged from 38 to 159 with a mean of 68 (1.7 ha).  
199 For 4-7 sample plots per stand, this resulted in sampling fractions of 4%-15% with a mean of  
200 10% and corresponding finite population corrections (eq. (4)) of 0.84-0.96 with a mean of 0.90.



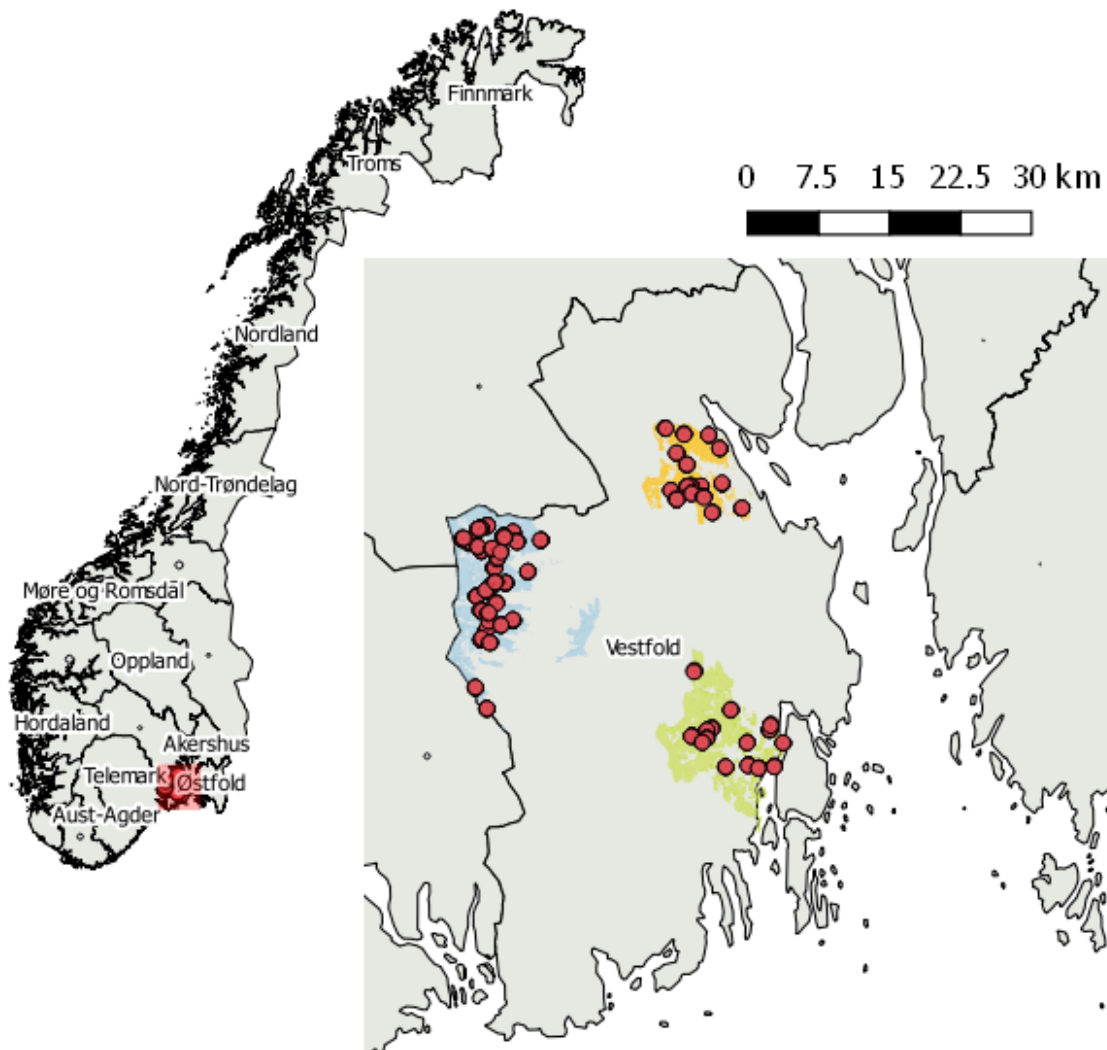


Figure 1: Left: Location of the study area within Norway (red square). Right: Location of the sample plots within the forest areas of Holmestrand (orange), Lardal (blue), and Stokke (green).

## 201 3.2. Modeling

### 202 3.2.1. Unit-level models

203 The unit-level variable of interest  $y_{ij}$  in the mixed effects model (6) was timber volume  
204 scaled to per-hectare values observed on the sample plots. Possible effects of unequal sampling  
205 fractions among the stands were ignored when estimating the model parameters. Based on  
206 previous experience, AP mean height observed on a sample plot ( $x_{2ij}$ ) and its square ( $x_{3ij}$ ) were  
207 the only explanatory variables in models with intercept ( $x_{1ij} = 1$ ), resulting in  $p = 3$ . Restricted  
208 maximum likelihood was used to estimate the model parameters. The fixed part of the mixed  
209 models explained more than 74% of the total variation.

210 Models ignoring heteroscedasticity ( $k_{ij} = 1$ , eq. 9) showed clear patterns of increasing variance  
211 with increasing  $x$  values in the residuals (Fig. 2 A and B). The constant  $k_{ij}$  for considering  
212 heteroscedasticity affects standard errors of estimates, and to a smaller degree, also the estimates  
213 themselves. Because influential observations may have a strong influence of the decision, we used  
214 the following procedure to select  $k_{ij}$ :

- 215 • The Akaike's information criterion (AIC) was minimized by varying the parameter  $\xi$  in  
216  $k_{ij} = x_{2ij}^{\xi}$ .
- 217 • Five influential observations were removed temporarily from the data set after inspecting  
218 the  $\sigma_v$ -residuals (eq. A.1) and  $\sigma_e$ -residuals (eq. A.2). The sample plot with the largest  
219 observed timber volume was among the influential observations.
- 220 • Minimizing the AIC with the reduced data set still left structures in the residuals. There-  
221 fore, the harmonic mean of the p-values of Breusch-Pagan tests (Breusch and Pagan, 1979)  
222 for the squared  $\sigma_v$ - and  $\sigma_e$ -residuals was maximized under the constraint that both p-values  
223 were  $> 0.05$  by varying the parameter  $\xi$  in  $k_{ij} = x_{2ij}^{\xi}$ . With p-values  $> 0.05$ , the hypothesis  
224 of homoscedasticity is not rejected given a 95% significance level.

225 The optimized value was  $\xi = 0.48$  resulting in Breusch-Pagan p-values of 0.14 and 0.10 for the  $\sigma_v$ -  
226 and  $\sigma_e$ -residuals, respectively. Breusch-Pagan p-values were smaller than 0.05 for the full data  
227 set with the influential observations. Larger values of  $\xi$  would have led to a stronger visibility of  
228 heteroscedasticity in the estimated standard errors but also to structures in the residuals. The  
229 values of  $k_{ij}$  ranged from 0.7 to 14.5 with a mean of 8.2. The value of  $k_{ij}$  is selected based on  
230 data although it is assumed to be known (Rao and Molina, 2015, Ch. 4.3). Possible influences  
231 on MSE estimators were ignored.

232 In some studies “outliers” are removed in order to meet the model assumptions (Battese et al.,  
233 1988). The four influential observations that were removed temporarily in order to select  $\xi$  were,  
234 however, not outliers because we could not find any evidence of data errors. We therefore kept all  
235 observations in the models but also report results for models without the influential observations.

236 Although the scaled residuals were approximately symmetrically distributed around zero  
237 (Fig. 2 C and D), it was not possible to select  $k_{ij}$  values that would have resulted in normally  
238 distributed residuals (test: Shapiro-Wilk given a 95% significance level), even after removing the  
239 influential observations.

240 For stand-level estimates, the stand-level mean of AP mean height observed on all grid  
241 cells ( $\bar{x}_{2iP}$ ), and the mean of squared AP mean height ( $\bar{x}_{3iP}$ ) were calculated. For considering  
242 small population sizes (eq. (10)), stand-level means need to be calculated also for the non-  
243 sampled population units ( $\bar{x}_{iR}$ , eq. (A.7)). However, the sample plots are circular and not  
244 aligned with the grid cells. Therefore, for each sample plot the closest grid cell was omitted in  
245 order to calculate  $\bar{x}_{2iR}$ , and  $\bar{x}_{3iR}$ . Table 1 summarizes some characteristics of the response and  
246 explanatory variables. Synthetic estimates (eq. (8)) were generated for the 64 stands with plots  
247 by assuming that no field data were available for them.

	Mean	Min	Max	SD
$y_{ij}$	193.02	0.00	947.80	141.23
$x_{2ij}$	91.70	0.47	263.73	57.11
$\bar{x}_{2iP}$	91.73	8.31	201.92	45.75

Table 1: Characteristics of plot-level timber volume ( $y_{ij}$ , m<sup>3</sup>/ha), plot-level AP mean height ( $x_{2ij}$ , dm), and stand-level means of AP mean height observed on all grid cells ( $\bar{x}_{2iP}$ , dm).

### 248 3.2.2. Area-level models

249 The response variable in the area-level model (eq. (11)) was the direct mean timber volume  
250 estimate of the sample plots ( $\hat{Y}_i^D$ , eq. (2)). The stand-level mean of AP mean height observed  
251 on all grid cells ( $\bar{x}_{2iP}$ ) was the only explanatory variable in models with intercept ( $\bar{x}_{1iP} = 1$ )  
252 such that  $p = 2$ . As opposed to the unit-level model, the parameter estimate of the square of  
253 the AP mean height was not significantly different from 0. Restricted maximum likelihood was  
254 used to estimate the model parameters.

255 The constant  $b_i$  for considering heteroscedasticity affects standard errors of estimates, and to  
256 a smaller degree, also the estimates themselves. Furthermore, influential observations may have  
257 a strong influence on the selection of  $b_i$ . Models ignoring heteroscedasticity ( $b_i = 1$ ) showed clear  
258 patterns of increasing variability in the residuals with increasing  $\bar{x}_{2iP}$  (Fig. 3 A). However, values

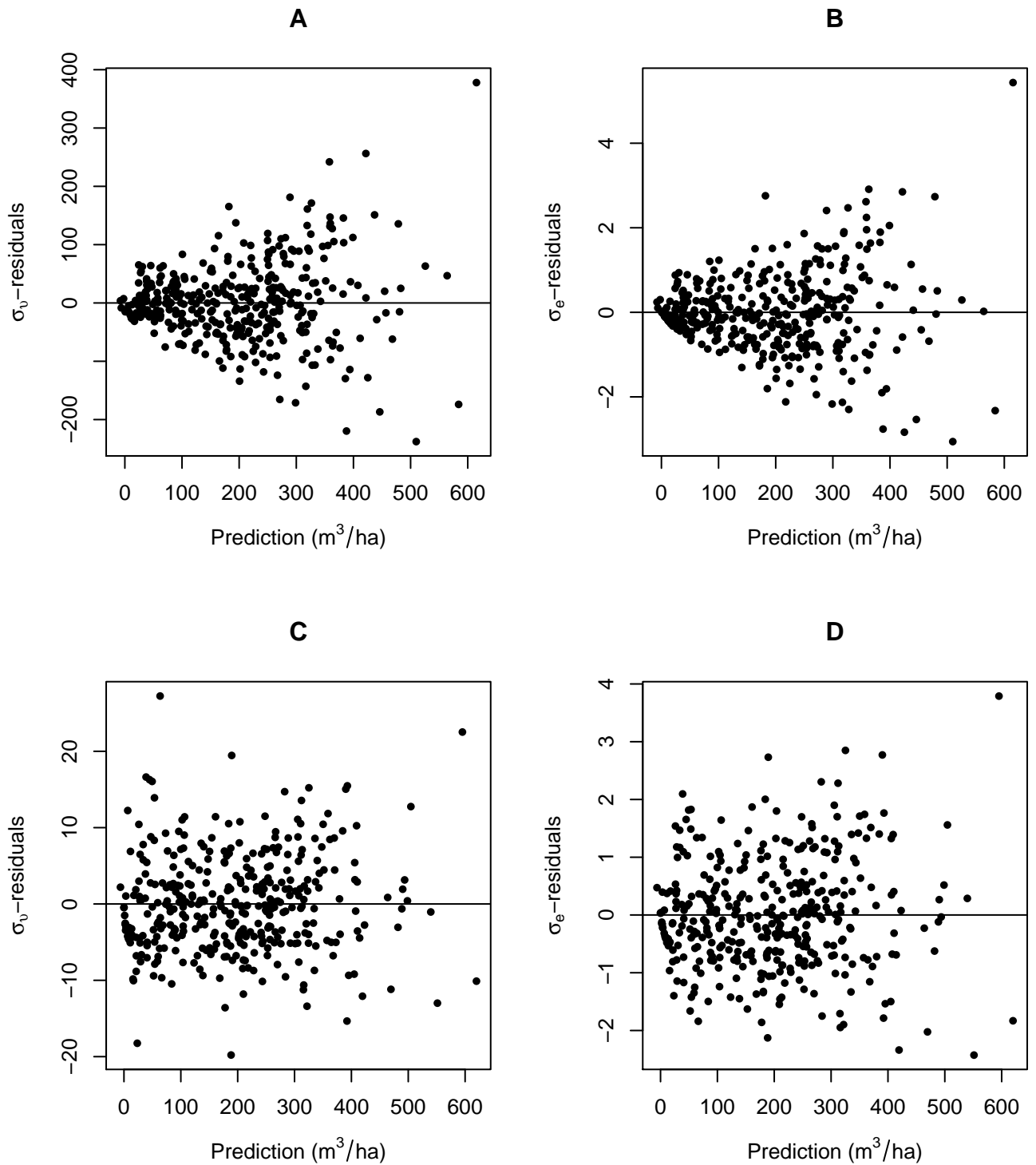


Figure 2: Transformed and scaled residuals (see AppendixA.1.1, including influential observations) versus predicted values for the unit-level model with  $k_{ij} = 1$  (A and B), and  $k_{ij} = x_{2ij}^{0.48}$  (C and D).

259 of  $b_i$  in the order of  $\bar{x}_{2iP}$  led to numeric instability. Therefore, we used the following procedure  
260 to select  $b_i$ :

- 261 • The explanatory variable  $\bar{x}_{2iP}$  was transformed ( $\bar{x}_{2iP}^*$ ) to the range  $0, \dots, 1$  and the AIC  
262 was minimized by varying the parameter  $\zeta$  in  $b_i = \bar{x}_{2iP}^* + \zeta$ .
- 263 • One influential observation was removed temporarily from the data set and scaled random  
264 effects (eq. A.15) were obtained. The influential observation was also among the influential  
265 observations in the unit-level model.
- 266 • Minimizing the AIC with the reduced data set still resulted in structures in the residuals.  
267 Therefore, the harmonic mean of the p-values of Breusch-Pagan and Shapiro-Wilk tests for  
268 the scaled random effects was maximized under the constrain that both p-values were  $>$   
269  $0.05$  by varying the parameter  $\zeta$  in  $b_i = \bar{x}_{2iP}^* + \zeta$ .

270 The selected value was  $\zeta = 0.39$  resulting in Breusch-Pagan and Shapiro-Wilk p-values of 0.32  
271 and 0.14, respectively. The Breusch-Pagan p-value was larger than 0.05 for the data set including  
272 the influential observation but the Shapiro-Wilk p-value was smaller than 0.05 (Fig. 3 B). As  
273 before, we did not have any indication of errors and therefore did not exclude the influential  
274 observation. The values of  $b_i$  ranged from 0.39 to 1.39 with a mean of 0.82. Possible influences  
275 on MSE estimators due to the selection of  $b_i$  were ignored.

### 276 3.3. Results

#### 277 3.3.1. Comparison of unit- and area-level estimators

278 This section gives an overview of the results which are further described in the two following  
279 sections. Due to the small sample size within stands, results of the design-based direct estimator  
280 are presented with the area-level estimates.

281 Although unit-level and area-level EBLUP estimates at stand-level were similar (Fig. 4),  
282 their SE differed considerably (Fig. 4 and 5). When not considering heteroscedasticity, SE of  
283 unit-level EBLUP estimates varied unrealistically little, and were on average similar to SE of  
284 area-level EBLUP estimates where SE increased with increasing estimates. Furthermore, SE of  
285 synthetic estimates were similar for unit-level and area-level models. However, not considering  
286 heteroscedasticity strongly violated the model assumption of homogeneity of variance.

287 In models where heteroscedasticity was addressed, SE of unit-level EBLUP estimates in-  
288 creased in parallel with increasing mean estimates but were on average 31% smaller than SE of

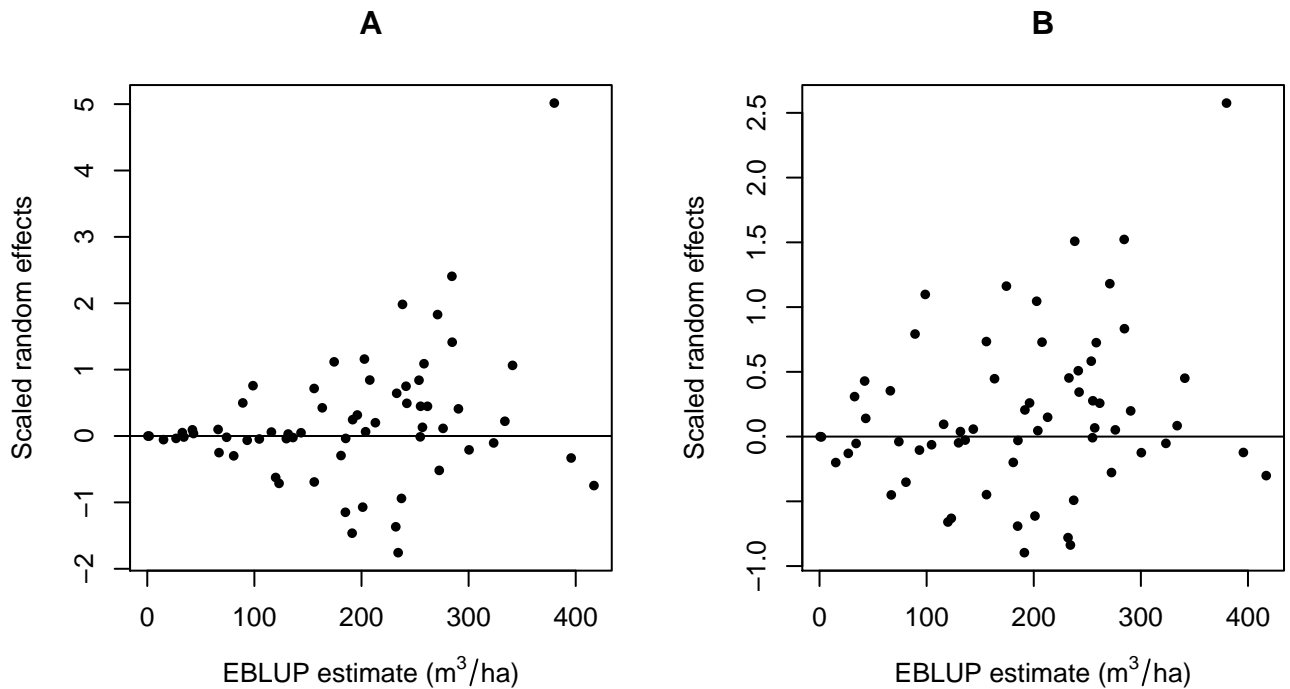


Figure 3: Scaled random effects (including one influential observation) versus area-level EBLUP estimates on stand-level with  $b_i = 1$  (A), and  $b_i = \bar{x}_{2iP} + 0.39$  (B).

289 area-level estimates. However, for 11 of the 64 stands, the SE of area-level EBLUP estimates was  
 290 slightly smaller than the SE of unit-level EBLUP estimates. These were typically stands with  
 291 relatively precise direct estimates. SE of unit-level synthetic estimates were consistently and, on  
 292 average, 28% smaller than SE of area-level synthetic estimates.

293 The influence of heteroscedasticity was hardly visible in SE unit-level and area-level synthetic  
 294 estimates. For unit-level estimates, this is because the contribution of the residual variance  
 295 ( $k_{ij}^2 \sigma_e^2$ ) that models the heteroscedasticity is negligible as it is divided by a large number of grid  
 296 cells (eq. A.11). Therefore, heteroscedasticity is only indirectly accounted for by the uncertainty  
 297 of the fixed parameter estimates. Also for area-level synthetic estimates, heteroscedasticity is  
 298 only accounted for by the uncertainty of the fixed parameter estimates (A.21).

### 299 3.3.2. Unit-level estimators

300 Stand-level synthetic estimates (assuming no sample plots within stands, eq. (8)) of mean  
 301 timber volume were similar whether heteroscedasticity was considered ( $k_{ij} = x_{2ij}^{0.48}$ ) or not ( $k_{ij} =$   
 302 1) and whether influential observations were included in the model or not (Fig. B.6 A). In other  
 303 words, the fixed effects estimates differed only slightly between the models. EBLUP estimates  
 304 ( $k_{ij} = x_{2ij}^{0.48}$ , all observations) of stand-level timber volume ranged from 6.13 to 437.77 with a

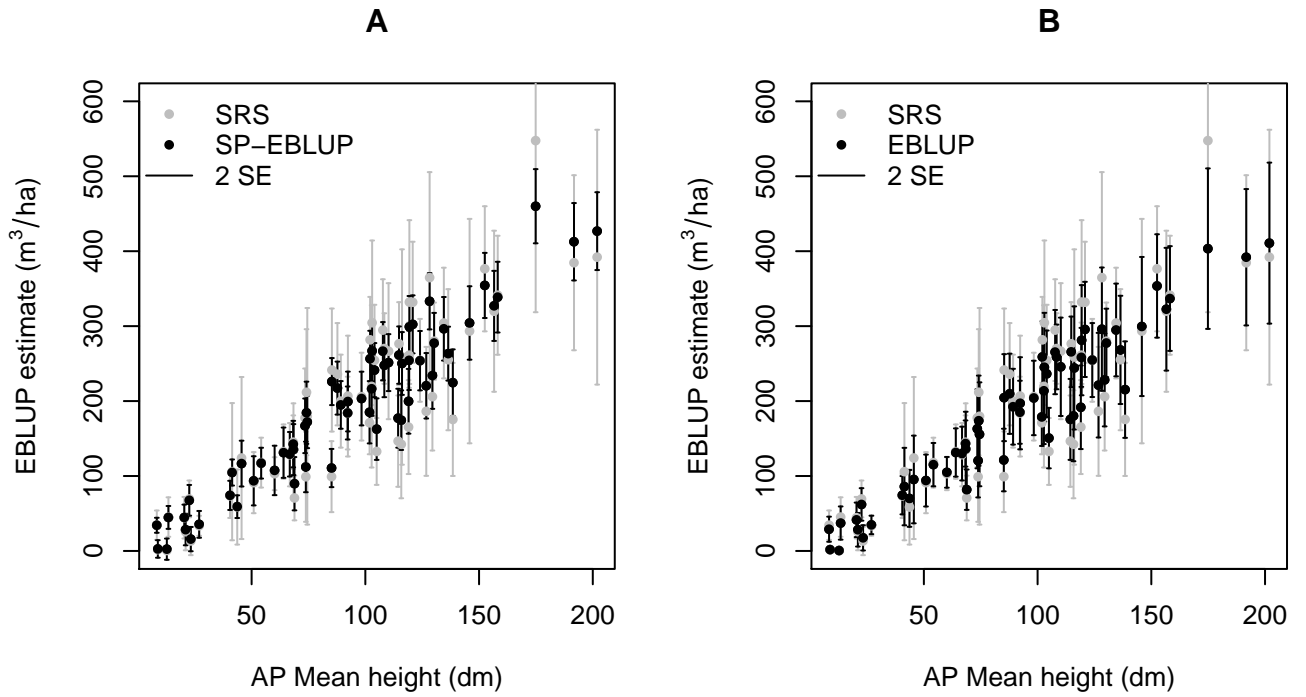


Figure 4: Stand-level estimates assuming small populations (SP-EBLUP) based on unit-level models (A) and area-level models (B). SRS = direct estimates.

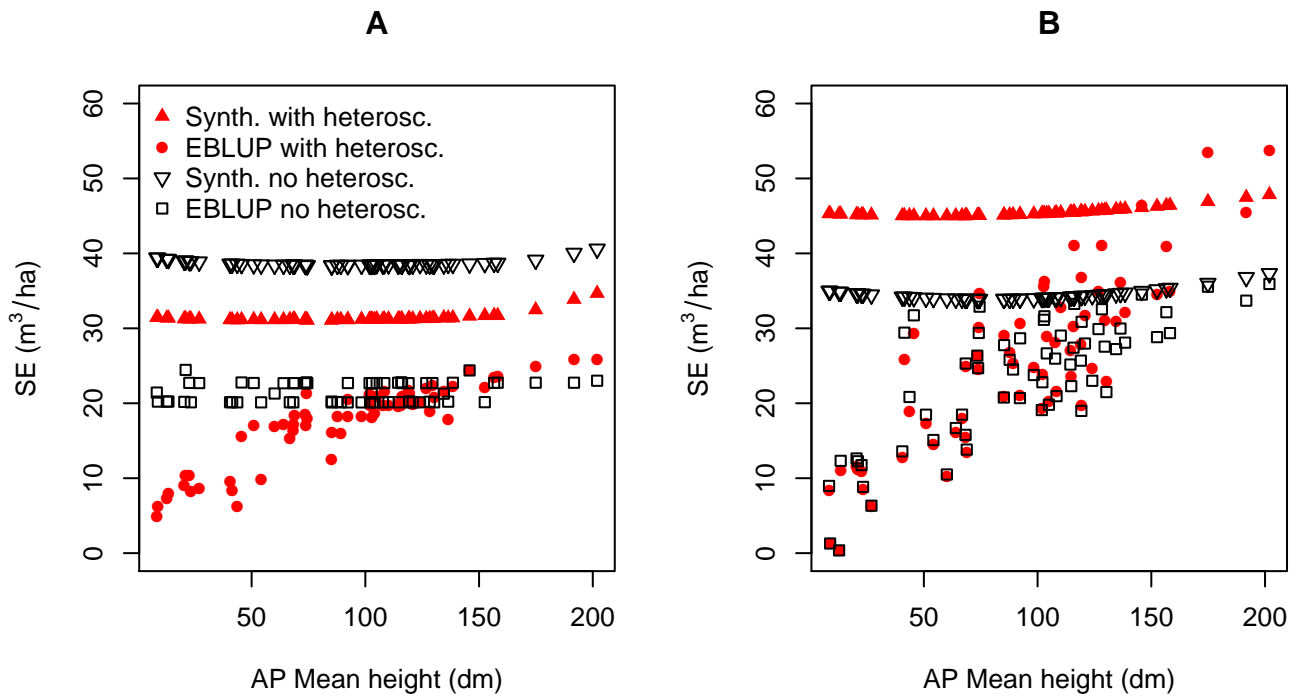


Figure 5: Standard errors versus AP mean height for unit-level (A) and area-level estimates (B).

305 mean of 190.83 m<sup>3</sup>/ha.

306 SE increased slightly for models including influential observations (Tab. B.7). The fact that  
307 mean estimates remained similar while standard errors increased made us more confident in  
308 keeping the influential observations despite of the formal violation of the assumptions of ho-  
309 moscedasticity and normality of the residuals.

310 SE of mean estimates based on models considering heteroscedasticity or not differed drasti-  
311 cally (Fig. 5 A and Tab. 3) which also resulted in differences in the EBLUP estimates (Fig. B.6  
312 B). Random-effect variances of models assuming heteroscedasticity were smaller than of models  
313 assuming homoscedasticity (Tab. 2), which resulted in larger values of  $\hat{\gamma}_i$  and thus more weight  
314 assigned to the synthetic estimator.

	$R^2$	$\sigma_v^2$	$\sigma_e^2$	mean( $\gamma_i$ )	min( $\gamma_i$ )	max( $\gamma_i$ )
k=1 all obs.	0.75	1437.71	3744.15	0.69	0.61	0.73
k=f(x) all obs.	0.74	943.55	44.86	0.67	0.38	0.98
k=1 sel. obs.	0.75	1401.29	3331.46	0.71	0.63	0.75
k=f(x) sel. obs.	0.75	995.00	41.08	0.69	0.38	0.98

Table 2: Unit-level model characteristics. k=1: no heteroscedasticity; k=f(x): heteroscedasticity considered; all obs.: all observations; sel. obs.: selected observations (without influential observations).

315 SE of EBLUP estimates assuming homoscedasticity are almost exclusively dependent on the  
316 number of observations within the stands (eqs. (9) and (A.4)), which made them appear rather  
317 unrealistic for the given data set. Most of the SE follow two imaginary lines in Fig. 5 A (black  
318 hollow squares) for stands with 7 sample plots (SE approximately 20 m<sup>3</sup>/ha) and 5 sample plots  
319 (SE approximately 23 m<sup>3</sup>/ha). Deviations from the two lines are visible for four stands with  
320 6 sample plots (SE between 20 and 23 m<sup>3</sup>/ha) or 4 sample plots (SE greater than 23 m<sup>3</sup>/ha).  
321 Some smaller deviations are caused by the secondary error terms (eqs. (A.5) and (A.6)). The  
322 influence of sample size on SE of EBLUP estimates was visible because the sampling fraction  
323 within stands was variable. SE of synthetic estimates (the same stands ignoring the sample plots)  
324 were, on average, almost twice as large as SE of EBLUP estimates (the same stands considering  
325 the sample plots) and, as expected, increased toward the extremes of the explanatory variable  
326 (Fig. 5 A, black hollow triangles).

327 SE of EBLUP estimates considering heteroscedasticity showed an increasing trend over the  
328 explanatory variable (Fig. 5 A, red filled dots) and were, on average, smaller than SE of EBLUP  
329 estimates assuming homoscedasticity. SE of synthetic estimates considering heteroscedasticity  
330 (Fig. 5 A, red filled triangles) exhibited a stronger increase toward the maximum than toward



331 the minimum of the explanatory variable. SE of synthetic estimates were always larger than SE  
332 for stands with observations but the difference decreased with increasing estimates (Fig. 5 A, red  
333 filled symbols).

	mean(SE)	min(SE)	max(SE)	mean(SE%)	min(SE%)	max(SE%)
k=1 all obs.	21.48	20.11	24.45	22.30	5.15	219.86
k=f(x) all obs.	17.42	4.92	25.89	13.55	5.55	100.74
k=1 all obs. synth.	38.64	38.35	40.59	35.09	8.91	199.87
k=f(x) all obs. synth.	31.38	31.08	34.67	25.43	7.51	110.30

Table 3: Standard errors (SE, m<sup>3</sup>/ha) and relative SE (%) of unit-level estimates. k=1: no heteroscedasticity; k=f(x): heteroscedasticity considered; all obs.: all observations; synth.: synthetic estimate.

334 Considering the small population size (eq. 10) had only minor effects for most stands (see  
335 AppendixB.2).

### 336 3.3.3. Area-level estimators

337 Direct (SRS) estimates of stand-level timber volume ( $\hat{Y}_i^D$ ) are a part of area-level EBLUP  
338 estimates (eq. 13), and ranged from 0.6 to 547.6 with a mean of 193.5 m<sup>3</sup>/ha. Standard errors  
339 (SE) including finite population correction (fpc, eq. 4) were on average 2 percentage points (pp)  
340 smaller than SE not considering fpc (Tab. 4).

	mean(SE)	min(SE)	max(SE)	mean(SE%)	min(SE%)	max(SE%)
SRS	34.35	0.38	114.47	22.37	8.33	81.33
SRS fpc	32.67	0.37	107.61	21.24	7.85	74.58

Table 4: Standard errors (SE) in m<sup>3</sup>/ha and SE relative to the estimate (%) of direct estimates (SRS) with and without finite population correction (fpc).

341 Differences between mean estimates and SE of area-level EBLUP estimates based on models  
342 including all observations or omitting one influential observation were less than 1%. Results  
343 for omitting one influential observation are therefore not reported. This suggests that a minor  
344 violation of model assumptions as indicated by test statistics (the null-hypothesis of normal dis-  
345 tributed residuals was rejected when including one influential observation) was of little practical  
346 relevance. Some model characteristics can be found in Table 5.

347 Considering heteroscedasticity ( $b_i = x_{2i}^* + 0.39$ ) or not ( $b_i = 1$ ), had little effect on the  
348 fixed-effects parameter estimates  $\hat{\beta}$  (Fig. B.8 A), however EBLUP estimates (Fig. B.8 B) and  
349 SE changed considerably (Fig. B.9, Tab. 6). EBLUP estimates considering heteroscedasticity  
350 ranged from 0.6 m<sup>3</sup>/ha to 411.0 m<sup>3</sup>/ha with a mean of 186.8 m<sup>3</sup>/ha.

351 Assuming heteroscedasticity in the model, resulted in almost a doubling of the random-effect  
352 variance (Fig. B.9, Tab. 5), and thus in more weight on the direct estimator (eq. (13)), and

353 consequently a larger SE of EBLUP estimates. This effect was most noticeable in stands with  
354 large SE (Fig. B.9 A). For synthetic estimates (assuming observations were not available), the  
355 SE increased on average by 12 pp (Fig. B.9, Tab. 6).

356 In tendency,  $\hat{\gamma}_i$  decreased with increasing EBLUP estimates which is the reason for similar  
357 SE of area-level estimates and direct (SRS) estimates up to direct SE of approximately 30 m<sup>3</sup>/ha  
358 (Fig. B.9 B). For the smallest mean estimate,  $\hat{\gamma}_i$  was close to 1 such that the EBLUP estimate  
359 was approximately equal to the direct (SRS) estimate (Tab. 5).

	$R^2$	$\sigma_v^2$	mean( $\gamma_i$ )	min( $\gamma_i$ )	max( $\gamma_i$ )
b=1	0.78	1116.54	0.57	0.09	1.00
b=f(x)	0.78	2002.78	0.61	0.19	1.00

Table 5: Area-level model characteristics for models including one influential observation. b=1: no heteroscedasticity; b=f(x): heteroscedasticity considered.

	mean(SE)	min(SE)	max(SE)	mean(SE%)	min(SE%)	max(SE%)
b=1 RV	23.10	0.37	35.92	17.31	7.39	74.05
b=f(x) RV	25.43	0.37	53.70	17.73	7.61	72.67
b=1 synth.	34.39	33.87	37.31	40.41	8.86	390.72
b=f(x) synth.	45.46	45.01	47.84	53.50	11.33	521.73

Table 6: Standard errors (SE, m<sup>3</sup>/ha) and SE relative to the estimate (%) of area-level estimates. b=1: no heteroscedasticity; b=f(x): heteroscedasticity considered; RV: variance of the direct estimate considered; synth.: synthetic estimate.

#### 360 4. Discussion

361 Individual tree crown approaches (ITC), area-based approaches (ABA), or domain-level ap-  
362 proaches (DLA) can utilize explanatory variables obtained from fine-resolution 3D remotely  
363 sensed data such as airborne laser scanning (ALS) or digital aerial photogrammetry (AP). The  
364 three approaches can be linked to two model types (unit-level and area-level) and their associ-  
365 ated model-based estimators. Unit-level and area-level EBLUP estimators are weighted averages  
366 of estimators that are exclusively dependent on the fitted linking model (synthetic estimators)  
367 and estimators that are based on the sample units within a domain. The weight depends on  
368 the model accuracy which again depends on the quality of the explanatory variables. In a case  
369 study, unit-level and area-level estimators were used for inference on mean timber volume within  
370 stands using linear linking models. Adjustments to the described methods would be required if  
371 non-linear models were to be applied (Rao and Molina, 2015, Ch. 4.6).

372 Satisfying all model assumptions (e.g., homogeneity of variance, normality of residual disper-  
373 sion) using real data can be difficult as we learned from our case study. However, small deviations  
374 from model assumptions may be inconsequential. For important decisions that require large ac-  
375 curacy, investing more in field sample plots deserves consideration. Due to their robustness  
376 against model-misspecification, model-assisted (design-based) estimators (e.g. Mandallaz, 2013;  
377 McRoberts et al., 2014) may be an interesting alternative to the model-based estimators used  
378 here if sufficient sample sizes per domain are available.

379 As observed in other studies (Hidiroglou and You, 2016; Mauro et al., 2017b), mean estimates  
380 from unit-level and area-level estimators can be similar. Furthermore, when satisfying the model  
381 assumptions unit-level estimators, on average, result in smaller standard errors (SE) than area-  
382 level estimators and SE decrease with domain size which is an intuitive property (Breidenbach  
383 et al., 2015). Furthermore, unit-level models can be used to generate sub-domain maps of forest  
384 resources. For these reasons, unit-level estimators may be preferred over area-level estimators, if  
385 the data afford their use.

386 However, the field data acquisition for area-level estimators can be considerably cheaper than  
387 for unit-level estimators, because exact sample locations are not required and efficient plot designs  
388 such as variable radius plots can be used without compromising the link to remotely sensed data.  
389 The reduced costs per sampling unit can be used toward a larger field sample which, in turn,  
390 may improve the precision of area-level estimates relative to unit-level estimates under a given  
391 budget.

392 Heteroscedasticity has a strong influence on the SE of unit-level and area-level estimates. If  
393 observed, it should be considered in the models to avoid violated assumptions and unrealistic SE.  
394 However, selecting the constants  $k_{ij}$  (unit-level) and  $b_i$  (area-level) for considering heteroscedas-  
395 ticity is a delicate matter because they affect standard errors of estimates and, to a smaller  
396 degree, also the estimates themselves. Therefore, we based the selection of these constants on  
397 objective methods that aim at fulfilling the model assumptions. Based on a visual inspection of  
398 residuals, Mauro et al. (2017b) selected similar values for the constants  $k_{ij}$  of unit-level models  
399 for estimating timber volume.

400 While the challenge of the chosen method for considering heteroscedasticity in unit-level  
401 models is to select an adequate value of the constant  $k_{ij}$ , the challenges of an alternative method  
402 proposed by Jiang and Nguyen (2012) is that the continuous nature of the heteroscedasticity  
403 is categorized and the number of categories has to be selected. Transformation of the response  
404 variable is another method that can help meeting model assumptions (e.g. Næsset, 1997). Infer-  
405 ence on domain means using transformed variables would, however, require modifications to the  
406 methods used here due to the required back-transformation (Rao and Molina, 2015, p. 140).

407 Considering heteroscedasticity in SAE deserves more attention. A reason for the limited  
408 attention to this topic in area-level estimators is that heteroscedasticity is naturally included  
409 because of their relation to the direct estimator. Furthermore, synthetic estimates using area-level  
410 estimators are often not considered when observations are available for each domain (Hidiroglou  
411 and You, 2016; Molina and Marhuenda, 2015). However, in forest inventories, the majority  
412 of stands may not contain any sample plots due to small sampling fractions and thus require  
413 synthetic estimates.

414 Remotely sensed data play a pivotal role in the discussed methods as they provide auxiliary  
415 information which are highly correlated to the variables of interest. Without the availability of  
416 auxiliary information, synthetic estimates (for stands without sample plots) would be of little  
417 practical relevance. With the availability of remotely sensed data that are less closely related  
418 to the variables of interest than the AP data we had available (for example Landsat images),  
419 synthetic estimates need to be used with care as they can have large systematic errors. Due to  
420 the advantages of coarser-resolution remotely sensed data, such as easier data handling, they may  
421 nonetheless be useful, especially for estimates on larger scales. With the availability of remotely  
422 sensed data that are more closely related to the variables of interest than the AP data we had  
423 available, issues of heteroscedasticity may be reduced.

424 To exploit remotely sensed data for estimation of volume and biomass we need auxiliary  
425 information that provide information proxies for stand structures related to densities of volume  
426 and biomass. Canopy height metrics and metrics related to canopy density, as provided by 3D  
427 remotely sensed data such as ALS and AP, are therefore key. But of course, environmental  
428 auxiliary information may also be useful, but typically more so the stronger the environmental  
429 gradients within the study region are expressed. For example, exposure and elevation can be  
430 useful predictors, but only when the vegetation is strongly influenced by these factors and if  
431 other auxiliary information (such as 3D remotely sensed data) have not already explained the  
432 variation in the vegetation. In sum, the scale of the study, and the study environment dictates  
433 the utility of the available auxiliaries.

## 434 5. Conclusions

435 The following conclusions can be drawn from this study. i) If present, including heteroscedas-  
436 ticity in models used for unit- and area-level SAE is important for obtaining realistic measures  
437 of precision. ii) SAE under heteroscedasticity should be studied more, especially for area-level  
438 estimation. The same is true for synthetic estimation in domains without samples, which is  
439 uncommon in area-level SAE. iii) On average, unit-level estimates can be expected to be more  
440 precise than area-level estimates. However, if the direct estimate (based on sample units only)  
441 has high precision (e.g., due to sufficient number of sample units and small variability such as in  
442 young stands), area-level estimates can have greater precision than unit-level estimates. iv) The  
443 use of digital aerial photogrammetry data considerably improved the precision of estimates.

## 444 6. Acknowledgments

445 This study was supported by the Norwegian Institute of Bioeconomy Research (NIBIO) and  
446 Nordic Forest Research (SNS) over the Centre of Advanced Research for the innovative use of 3D  
447 remote sensing in mapping of forest and landscape attributes based on national forest inventories  
448 (CARISMA). We would like to thank the editor and three anonymous reviewers for constructive  
449 comments.

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561 **Appendix A. Methods**

562 *Appendix A.1. Unit-level EBLUP estimators*

563 *Appendix A.1.1. Transformed and scaled residuals*

564 The transformed residuals ( $\sigma_v$ -residuals) are given by

$$u_{ij} = \frac{(y_{ij} - \hat{\tau} \hat{Y}_i^D) - (\mathbf{x}_{ij} - \hat{\tau} \bar{\mathbf{x}}_i)^T \hat{\boldsymbol{\beta}}}{k_{ij}} \quad (\text{A.1})$$

565 with  $\hat{\tau} = 1 - \sqrt{1 - \hat{\gamma}}$  where  $\hat{\gamma} = \frac{\hat{\sigma}_v^2}{\hat{\sigma}_v^2 + \hat{\sigma}_e^2}$  and  $\bar{\mathbf{x}}_i$  is the sample mean of the explanatory variables  
 566 (Militino et al., 2006). The standardized residuals ( $\sigma_e$ -residuals) are given by

$$\varepsilon_{ij} = \frac{e_{ij}}{k_{ij} \hat{\sigma}_e} \quad (\text{A.2})$$

567 where  $e_{ij}$  is an empirical residual.

568 *Appendix A.1.2. Estimating uncertainty*

569 For large populations (Rao and Molina, 2015, Ch. 7.2), the (unconditional) MSE is estimated  
 570 as the sum of three terms

$$\widehat{\text{MSE}}(\hat{\mu}_i^{\text{UE}}) = g_{1i} + g_{2i} + 2g_{3i} \quad (\text{A.3})$$

571 where

$$g_{1i} = (1 - \hat{\gamma}_i) \hat{\sigma}_v^2 \quad (\text{A.4})$$

572 describes the influence of the random-effect variance (Militino et al., 2007). Because  $g_{1i}$  is the  
 573 leading term (has the most influence on the MSE) and due to the structure of  $\hat{\gamma}_i$  (eq. 9), the  
 574 variability of the MSE among domains is almost exclusively affected by the number of samples  
 575 within a domain ( $n_i$ ) in the case of homoscedasticity ( $k_{ij} = 1$ ).

576 The term  $g_{2i}$  describes the uncertainty due to the estimation of the fixed-effects parameters

577  $\boldsymbol{\beta}$

$$g_{2i} = (\bar{\mathbf{x}}_{iP} - \hat{\gamma}_i \bar{\mathbf{x}}_{ia})^T \widehat{\text{Cov}}(\hat{\boldsymbol{\beta}}) (\bar{\mathbf{x}}_{iP} - \hat{\gamma}_i \bar{\mathbf{x}}_{ia}) \quad (\text{A.5})$$

578 where  $\bar{\mathbf{x}}_{ia} = \sum_{j=1}^{n_i} a_{ij} \mathbf{x}_{ij} / a_i$  is the weighted sample mean of the explanatory variables and  
 579  $\widehat{\text{Cov}}(\hat{\boldsymbol{\beta}})$  is the covariance matrix of the estimated fixed-effect parameters  $\hat{\boldsymbol{\beta}}$ .

580 The term  $g_{3i}$  describes the uncertainty due to the estimation of  $\sigma_v^2$  and  $\sigma_e^2$

$$g_{3i} = \frac{\hat{\sigma}_e^4 \bar{V}_v + \hat{\sigma}_v^4 \bar{V}_e - 2\hat{\sigma}_e^2 \hat{\sigma}_v^2 \bar{V}_{ve}}{a_i^2 (\hat{\sigma}_v^2 + \hat{\sigma}_e^2 a_i)^3} \quad (\text{A.6})$$

581 where  $\bar{V}_v, \bar{V}_e$  and  $\bar{V}_{ve}$  are the asymptotic variance and covariance estimates of  $\hat{\sigma}_v^2$  and  $\hat{\sigma}_e^2$ ,  
 582 respectively.

583 For small populations (Rao and Molina, 2015, Ch. 7.2.3), the MSE is estimated by

$$\widehat{\text{MSE}}(\hat{Y}_i^{\text{UE}}) = (1 - f_i)^2 \widetilde{\text{MSE}}(\hat{\mu}_i^{\text{UE}}) + g_{4i} \quad (\text{A.7})$$

584 where

$$\widetilde{\text{MSE}}(\hat{\mu}_i^{\text{UE}}) = g_{1i} + \tilde{g}_{2i} + 2g_{3i}. \quad (\text{A.8})$$

585  $\tilde{g}_{2i}$  is obtained by substituting the population mean of the explanatory variables  $\bar{\mathbf{x}}_{iP}$  with the  
 586 mean of the the explanatory variables of the non-sampled population units  $\bar{\mathbf{x}}_{iR}$  in  $g_{2i}$  (eq. A.5).

587 The fourth component of the MSE estimate considers the residual error variance

$$g_{4i} = g_{4i}^* \hat{\sigma}_e^2 \quad (\text{A.9})$$

588 where

$$g_{4i}^* = \frac{\mathbf{k}_{iR}^T \mathbf{k}_{iR}}{N_i^2} \quad (\text{A.10})$$

589 is the proportion to which the residual error variance is incorporated into the MSE, with  $\mathbf{k}_{iR}$  as  
 590 the vector of constants for considering heteroscedasticity of the non-sampled population units.

591 Under homoscedasticity, the term  $g_{4i}^*$  reduces to  $g_{4i}^* = N_i^{-2}(N_i - n_i)$  which means that it is not  
 592 necessary to know which population unit is part of the sample in that case. The term  $g_{4i}$  can  
 593 also be used to consider spatial autocorrelation (Mauro et al., 2017a; Breidenbach et al., 2015),  
 594 which is, however, outside the scope of this study.

595 The MSE of the synthetic estimator (eq. 8) results from

$$\widehat{\text{MSE}}(\hat{\mu}_i^{\text{US}}) = g_{1i} + g_{2i} + g_{4i} \quad (\text{A.11})$$

596 by setting  $\hat{\gamma}_i := 0$  and noting that  $\mathbf{k}_{iR} = \mathbf{k}_{iP}$  and  $n_i = 0$  in  $g_{4i}$  (eq. (A.9)). Under homoscedas-

597 ticity,  $g_{4i}$  reduces to

$$g_{4i} = \frac{\hat{\sigma}_e^2}{N_i} \quad (\text{A.12})$$

598 (McRoberts, 2006; Breidenbach et al., 2015). From this form of the component  $g_{4i}$  it becomes  
 599 clear how quickly the influence of the residual error variance reduces with domain size and that  
 600  $g_{4i}$  can be ignored for large populations.

601 *Appendix A.2. Area-level EBLUP estimators*

602 *Appendix A.2.1. Estimating model parameters*

603 The fixed model parameters are estimated by

$$\hat{\beta} = \left( \sum_{i=1}^m \frac{\bar{\mathbf{x}}_{iP} \bar{\mathbf{x}}_{iP}^T}{\hat{\vartheta}} \right)^{-1} \left( \sum_{i=1}^m \frac{\bar{\mathbf{x}}_{iP} \hat{Y}_i^D}{\hat{\vartheta}} \right) \quad (\text{A.13})$$

604 where  $\hat{\vartheta} = \hat{\sigma}_i^2 + \hat{\sigma}_v^2 b_i^2$  is the estimated total model variance (Rao and Molina, 2015, Ch. 6.1.1).

605 The random effects are indirectly estimated by

$$\hat{v}_i = \frac{\hat{Y}_i^D - \hat{\mu}_i^{\text{AE}}}{b_i} \quad (\text{A.14})$$

606 and scaled random effects result from

$$\hat{\varphi}_i = \frac{\hat{v}_i}{\hat{\sigma}_v^2}. \quad (\text{A.15})$$

607 *Appendix A.2.2. Estimating uncertainty*

608 The MSE of area-level estimates as described by Fay and Herriot (1979) can be estimated by  
 609 adding the terms

$$\widehat{\text{MSE}}(\tilde{\mu}_i^{\text{AE}}) = g_{1i} + g_{2i} + 2g_{3i} \quad (\text{A.16})$$

610 where

$$g_{1i} = (1 - \hat{\gamma}_i) \hat{\sigma}_v^2 \quad (\text{A.17})$$

611 reflects the uncertainty in the random effect,

$$g_{2i} = (1 - \hat{\gamma}_i)^2 \bar{\mathbf{x}}_{iP}^T \widehat{\text{Cov}}(\hat{\beta}) \bar{\mathbf{x}}_{iP} \quad (\text{A.18})$$

612 reflects the uncertainty in the fixed model parameter estimates  $\hat{\beta}$ , and

$$g_{3i} = \hat{\sigma}_i^4 (\hat{\sigma}_i^2 + \hat{\sigma}_v^2)^{-3} \bar{V}_v \quad (\text{A.19})$$

613 reflects the uncertainty due to estimating the random effect, where  $\bar{V}_v$  is the asymptotic variance  
 614 of the random effect.

615 The assumptions in the MSE estimator (A.16) include that the variance of the direct estimate  
 616 is known without uncertainty. In practice, the variance  $\hat{\sigma}_i^2$  is estimated (i.e., is not known without  
 617 uncertainty) and is either directly plugged-in to (A.16) as described in the equations above or  
 618 after smoothing. However, Wang and Fuller (2003) described general estimators for area-level  
 619 models where the assumptions about known model parameters are relaxed. Similarly, Rivest  
 620 and Vandal (2003) developed estimators for the special case where the direct estimate is based  
 621 on unit-level samples as in our case study. To account for the uncertainty in the variance of the  
 622 direct estimate  $\hat{\sigma}_i^2$ , an additional term is added to the MSE estimator (A.16)

$$\widehat{\text{MSE}}_{RV}(\hat{\mu}_i^{\text{AE}}) = \widehat{\text{MSE}}(\hat{\mu}_i^{\text{AE}}) + 2(\hat{\sigma}_v^2 + \hat{\sigma}_i^2)^{-3} \hat{\sigma}_v^4 \hat{\delta} \quad (\text{A.20})$$

623 where  $\hat{\delta} = 2\hat{\sigma}_i^4/(n_i - 1)$ .

624 For domains without samples, the variance of the synthetic estimate (Rao and Molina, 2015,  
 625 Ch. 6.2.2) is the sum of the random effect variance estimate and the uncertainty in the model  
 626 parameter estimates

$$\widehat{\text{MSE}}(\tilde{\mu}_i^{\text{AS}}) = g_{1i} + g_{2i} \quad (\text{A.21})$$

627 by setting  $\hat{\gamma}_i := 0$ .

## 628 **AppendixB. Results**

### 629 *AppendixB.1. Unit-level EBLUP estimates*

### 630 *AppendixB.2. Unit-level EBLUP estimates for small populations*

631 For small populations, the samples have to be considered in the EBLUP estimates (eq. 10).  
 632 The latter were therefore more variable (Fig. B.7 A) than estimates assuming large populations.  
 633 EBLUP estimates of stand-level timber volume ranged from 3.35 to 440.89 with a mean of 190.35  
 634  $\text{m}^3/\text{ha}$ . Furthermore, the residual error term has to be considered in the standard errors (eq. A.7)  
 635 which had different effects for models considering heteroscedasticity or not (Fig. B.7 B and  
 636 Tab. 3). For models assuming homoscedasticity, the standard errors were generally smaller when  
 637 considering the sampling fraction. For models considering heteroscedasticity, the standard errors  
 638 for small populations were often smaller for  $x_{2ij} < 100$  dm, but in tendency larger otherwise. The  
 639 reason for this difference is that the residual error term receives more weight for larger estimates

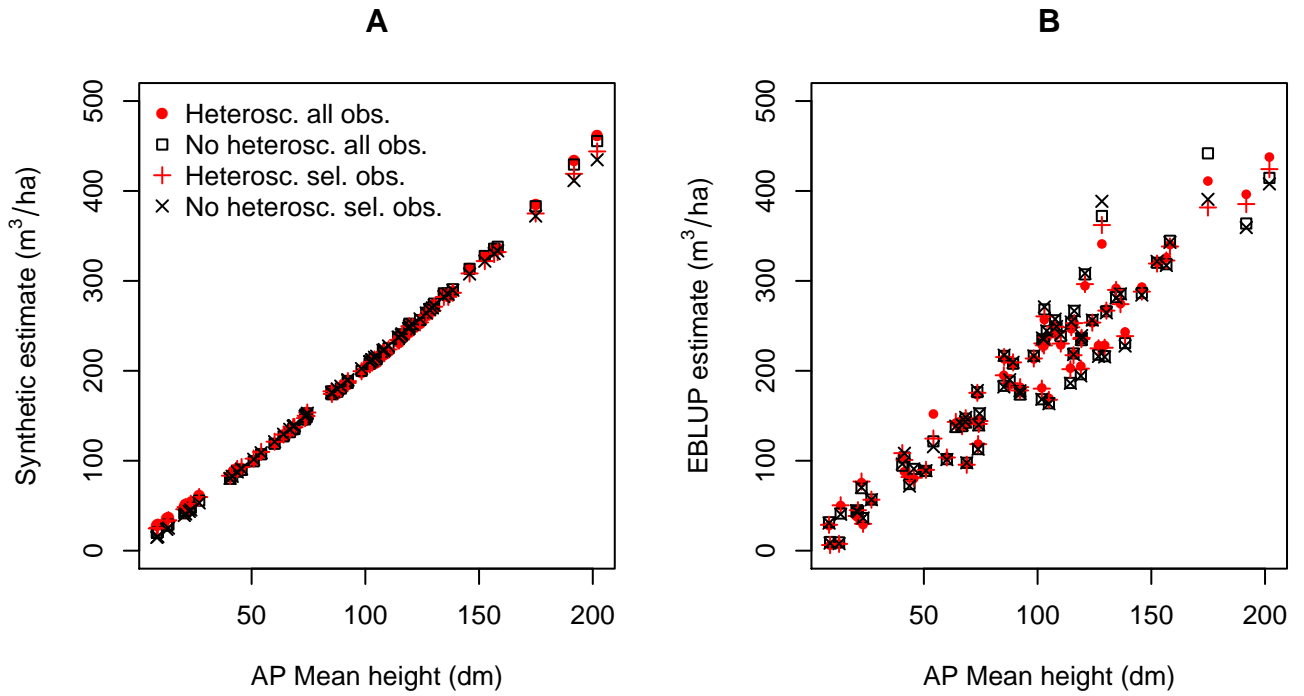


Figure B.6: Mean timber volume estimates using unit-level models for 64 stands. Synthetic estimates (A) using all observations or selected observations (omitting influential observations). EBLUP estimates (B).

640 due to heteroscedasticity which counter-acts the reduction of the SE by considering the sampling  
 641 fraction (eq. (A.7)).

	mean(SE)	min(SE)	max(SE)	mean(SE%)	min(SE%)	max(SE%)
k=1 all obs.	21.48	20.11	24.45	22.30	5.15	219.86
k=f(x) all obs.	17.42	4.92	25.89	13.55	5.55	100.74
k=1 sel. obs.	20.61	19.17	23.43	22.45	5.24	247.70
k=f(x) sel. obs.	17.29	4.72	25.84	13.28	5.55	95.35
k=1 all obs. synth.	38.64	38.35	40.59	35.09	8.91	199.87
k=f(x) all obs. synth.	31.38	31.08	34.67	25.43	7.51	110.30
SP k=1 all obs.	20.79	18.95	24.22	21.49	4.96	214.06
SP k=f(x) all obs.	17.18	5.01	25.96	13.30	5.50	95.74

Table B.7: Standard errors (SE, m<sup>3</sup>/ha) and relative SE (%) of unit-level estimates. k=1: no heteroscedasticity; k=f(x): heteroscedasticity considered; all obs.: all observations; sel. obs.: selected observations (without influential observations); synth.: synthetic estimate; SP: considering small population size.

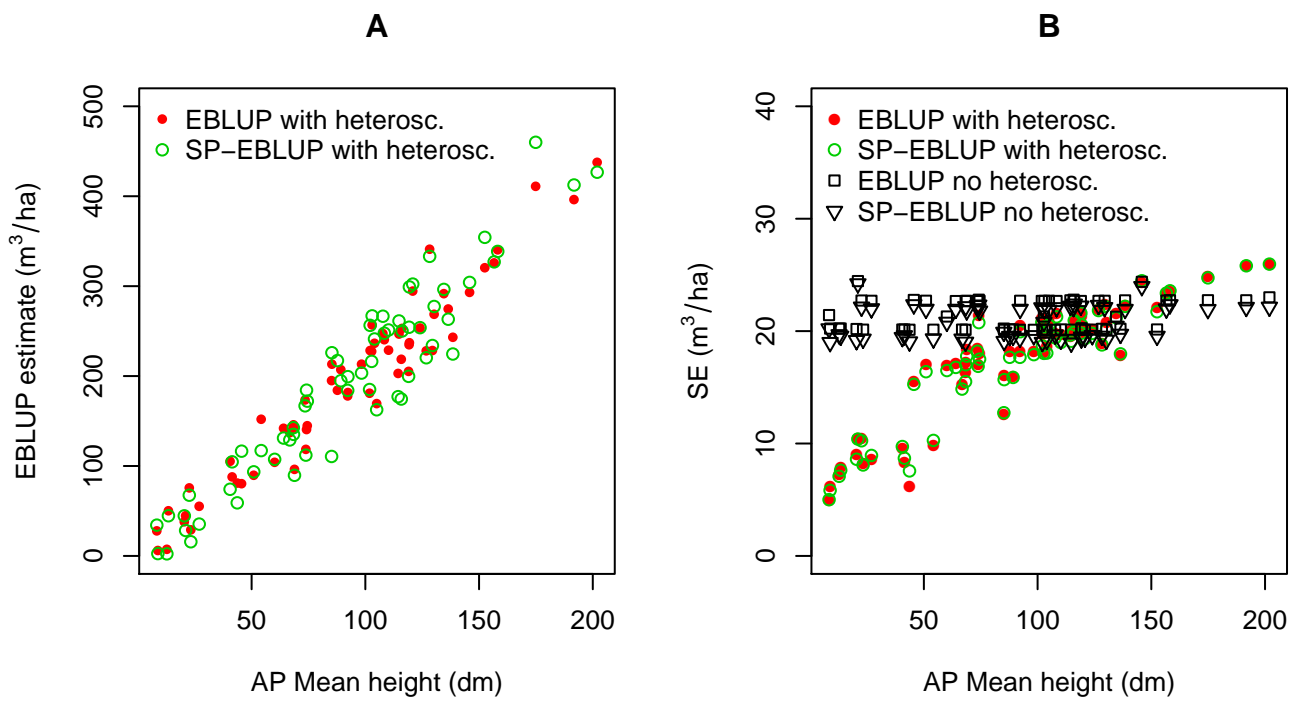


Figure B.7: EBLUP estimates assuming small populations (SP-EBLUP) and EBLUP estimates (A). Standard errors of EBLUP estimates (B).

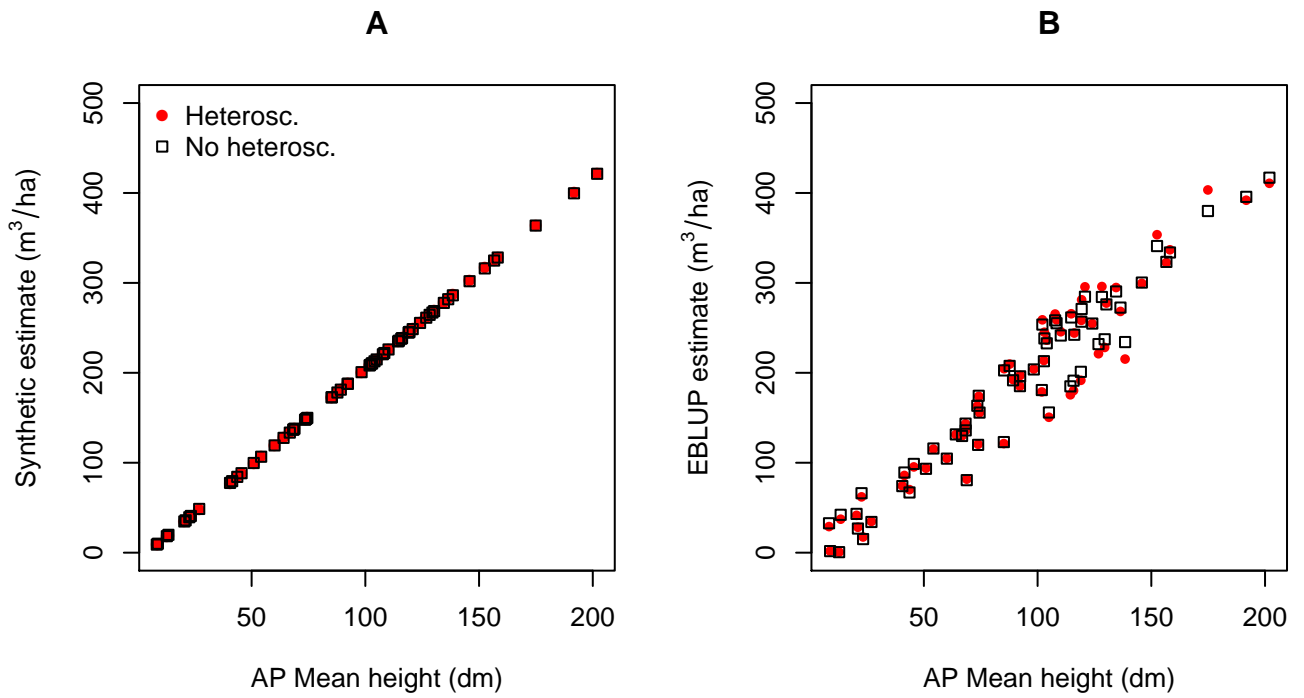


Figure B.8: Mean timber volume estimates using area-level models for 64 stands. Synthetic estimates (A), EBLUP estimates (B).

642 *Appendix B.3. Area-level estimates*

643 SE of estimates considering the variance of the direct estimator ( $\widehat{MSE}_{RV}(\hat{\mu}_i^{AE})$ ) were similar  
 644 but slightly larger than SE not considering the variance of the direct estimator ( $\widehat{MSE}_{FH}(\hat{\mu}_i^{AE})$ )  
 645 (Tab. B.8).

	mean(SE)	min(SE)	max(SE)	mean(SE%)	min(SE%)	max(SE%)
b=1 FH	21.44	0.37	34.38	16.33	6.96	74.01
b=f(x) FH	23.45	0.37	51.48	16.62	7.28	72.65
b=1 RV	23.10	0.37	35.92	17.31	7.39	74.05
b=f(x) RV	25.43	0.37	53.70	17.73	7.61	72.67
b=1 synth.	34.39	33.87	37.31	40.41	8.86	390.72
b=f(x) synth.	45.46	45.01	47.84	53.50	11.33	521.73

Table B.8: Standard errors (SE, m<sup>3</sup>/ha) and SE relative to the estimate (%) of area-level models. b=1: no heteroscedasticity; b=f(x): heteroscedasticity considered; FH: variance of the direct estimate not considered; RV: variance of the direct estimate considered; synth.: synthetic estimate.



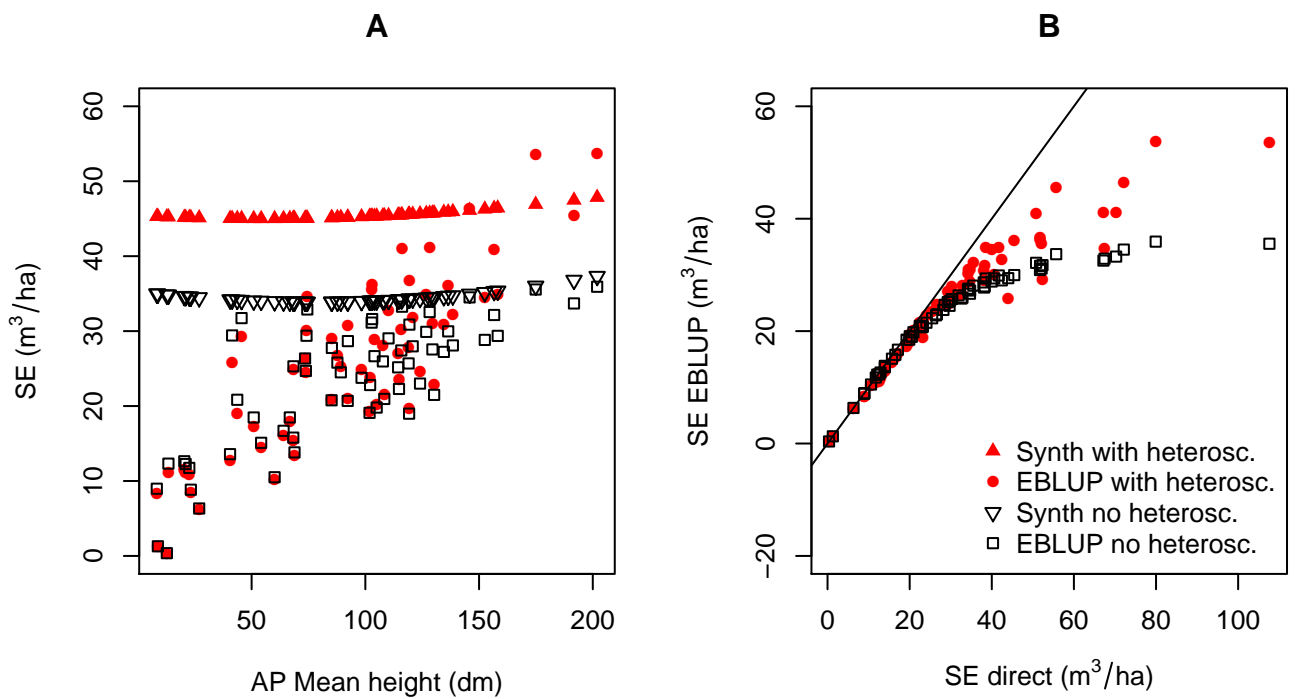


Figure B.9: Standard errors (SE) versus AP mean height for area-level estimates including all observations (A) and SE of EBLUP estimates vs SE of direct estimates (B).

646 LIST OF FIGURE CAPTIONS

647 **Figure 1** Left: Location of the study area within Norway (red square). Right: Location of the  
648 sample plots within the forest areas of Holmestrand (orange), Lardal (blue), and Stokke  
649 (green).

650 **Figure 2** Transformed and scaled residuals (see AppendixA.1.1, including influential obser-  
651 vations) versus predicted values for the unit-level model with  $k_{ij} = 1$  (A and B), and  
652  $k_{ij} = x_{2ij}^{0.48}$  (C and D).

653 **Figure 3** Scaled random effects (including one influential observation) versus area-level EBLUP  
654 estimates on stand-level with  $b_i = 1$  (A), and  $b_i = \bar{x}_{2iP} + 0.39$  (B).

655 **Figure 4** Stand-level estimates assuming small populations (SP-EBLUP) based on unit-level  
656 models (A) and area-level models (B). SRS = direct estimates.

657 **Figure 5** Standard errors versus AP mean height for unit-level (A) and area-level estimates (B).

658 **Figure B.6** Mean timber volume estimates using unit-level models for 64 stands. Synthetic  
659 estimates (A) using all observations or selected observations (omitting influential observa-  
660 tions). EBLUP estimates (B).

661 **Figure B.7** EBLUP estimates assuming small populations (SP-EBLUP) and EBLUP estimates  
662 (A). Standard errors of EBLUP estimates (B).

663 **Figure B.8** Mean timber volume estimates using area-level models for 64 stands. Synthetic  
664 estimates (A), EBLUP estimates (B).

665 **Figure B.9** Standard errors (SE) versus AP mean height for area-level estimates including all  
666 observations (A) and SE of EBLUP estimates vs SE of direct estimates (B).