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A Basic Model to Predict Enteric Methane Emission from Dairy Cows and Its Application to Update Operational Models for the National Inventory in Norway

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Simple Summary: Many techniques exist to quantify enteric methane (CH₄) emissions from dairy cows. Since measurement on the entire national cow populations is not possible, it is necessary to use estimates for national inventory reporting. This study aimed to develop (1) a basic equation of enteric CH₄ emissions from individual animals based on feed intake and nutrient contents of the diet, and (2) to update the operational way of calculation used in the Norwegian National Inventory Report based on milk yield and concentrate share of the diet. An international database containing recently published data was used for this updating process. By this the accuracy of the CH₄ production estimates included in the national inventory was improved.

Abstract: The aim of this study was to develop a basic model to predict enteric methane emission from dairy cows and to update operational calculations for the national inventory in Norway. Development of basic models utilized information that is available only from feeding experiments. Basic models were developed using a database with 63 treatment means from 19 studies and were evaluated against an external database ($n = 36$, from 10 studies) along with other extant models. In total, the basic model database included 99 treatment means from 29 studies with records for enteric CH₄ production (MJ/day), dry matter intake (DMI) and dietary nutrient composition. When evaluated by low root mean square prediction errors and high concordance correlation coefficients, the developed basic models that included DMI, dietary concentrations of fatty acids and neutral detergent fiber performed slightly better in predicting CH₄ emissions than extant models. In order to propose country-specific values for the CH₄ conversion factor Y_m (% of gross energy intake partitioned into CH₄) and thus to be able to carry out the national inventory for Norway, the existing operational model was updated for the prediction of Y_m over a wide range of feeding situations. A simulated operational database containing CH₄ production (predicted by the basic model), feed intake and composition, Y_m and gross energy intake (GEI), in addition to the predictor variables energy corrected milk yield and dietary concentrate share were used to develop an operational model. Input values of Y_m were updated based on the results from the basic models. The predicted Y_m ranged from 6.22 to 6.72%. In conclusion, the prediction accuracy of CH₄ production from dairy cows was improved with the help of newly published data, which enabled an update of the operational model for calculating the national inventory of CH₄ in Norway.

Keywords: dairy cattle; prediction model; methane conversion factor; dry matter intake; fatty acid; neutral detergent fiber

1. Introduction

The increase in global average surface temperature over the past half-century cannot be fully explained by natural climate variability. Scientific evidence indicates that the leading cause of climate change in the most recent half century is anthropogenic. Especially damaging is the increase in the concentration of atmospheric greenhouse gases (GHG), including carbon dioxide (CO₂), chlorofluorocarbons (CFCs), methane (CH₄), tropospheric ozone and nitrous oxide (N₂O) [1]. Animal husbandry is a source of anthropogenic GHG emission with CH₄ and N₂O as main gases, accounting for 30% of the total emissions by the agricultural sector [2]. Through CH₄, dairy production systems contribute, expressed in CO₂-equivalents, approximately one-half of the GHG emissions attributed to animal husbandry. Of this, on average 81% originate from enteric fermentation and 19% from manure [3]. Enteric CH₄ arises mainly as a side-product from rumen microbial fermentation of feed, especially fiber, to volatile fatty acids (VFAs). This fermentation process generates an excess of hydrogen (H₂) that is removed in the rumen by methanogens through reduction of CO₂ to CH₄.

The factors determining the amount of enteric CH₄ produced per animal include feed dry matter intake, diet composition (e.g., contents of ether extract (EE) or fatty acids (FAs) and neutral detergent fiber (NDF)), rumen microbial population, host physiology and host genetics [4]. To identify efficient mitigation strategies, the amount of CH₄ produced by the dairy system needs to be quantified as accurately as possible. Direct measurements of enteric CH₄ production (MJ/day) from cattle can be conducted using various methods, such as respiration chambers, sulfur hexafluoride (SF₆) tracer technique and the GreenFeed (GF) system (C-Lock Inc., Rapid City, SD, USA; [5]). However, when the total national CH₄ emissions need to be assessed for an inventory these techniques are not feasible due to the sheer number of measurements which would be needed. For this purpose, often quantitative approaches such as empirical modelling have been used to estimate CH₄ production in dairy cows [6,7].

Accurate information about feed intake and dietary composition is required for good prediction but this information is available only from feeding experiments and thus for a limited number of animals, while information about milk yield and dietary concentrate share is available for the Norwegian dairy cow population from the Dairy Herd Recording System (TINE SA, Oslo, Norway) for a continuous time series starting in 1990 [8]. Thus, the present study involved the development of an accurate basic model for prediction of enteric CH₄ production, and operational models for prediction of the CH₄ conversion factor (Y_m , % of gross energy intake (GEI) lost as CH₄). The Y_m is globally used for national GHG emission inventories and research on mitigation strategies [9]. Previously, Nielsen et al. [6] published in 2013 a basic model for the prediction of enteric CH₄ emission from dairy cows based on 47 treatment means from 12 studies. This equation is used in the Nordic Feed Evaluation System—NorFor [8]. One year later, Storlien et al. [7] developed another basic model based on 78 treatment means from 21 studies. This later model [7], and an operational model [8] using information about milk yield and concentrate share, are those which were used by the Norwegian Environment Agency (Miljødirektoratet) for the National Inventory Report to the United Nations Framework Convention on Climate Change (UNFCCC) and Kyoto Protocol/Paris Agreement. The operational model is dependent on the output of CH₄ production predicted by the basic model. The basic model [7] was developed based only on studies published until 2013. In addition, this model did not take into account the effect of dietary NDF.

Therefore the objectives of the present study were (1) to extend the database of Storlien et al. [7] with more recent studies; (2) to develop basic models using this extended database, and evaluate them against extant models in their performance in predicting enteric CH₄ production; (3) to use our best performing basic model to predict CH₄ production and to calculate Y_m with the help of the NorFor feed analysis database (NorFor-database) [8]; and (4) to update operational models where energy-corrected milk (ECM) and dietary concentrate share in the diet were used to predict Y_m and GEI, respectively.

2. Materials and Methods

The basic models were developed using information of CH₄ production, dry matter intake (DMI) and dietary nutrient compositions, from published feeding experiments. The operational model was developed to predict Y_m using energy corrected milk and dietary concentrate share based on an operational database (NorFor) [8] simulated to cover a wide range of feeding situations reported in the Dairy Herd Recording System (TINE SA, Oslo, Norway).

2.1. Basic Model Database

The basic model database originally used by Storlien et al. [7] was collated from 21 studies (Nordic, European, intercontinental) published from 1997 to 2013, consisting of 78 treatment means. The database was divided into two subsets, one for model development ($n = 42$) and one for model evaluation ($n = 36$). In the present study, the subset for basic model development from Storlien et al. [7] was extended by adding data published since 2013 where CH₄ production, forage proportion, DMI and contents of EE or FAs and NDF in diets for dairy cows were reported ($n = 21$ treatment means from 8 studies, highlighted in grey shading in Table 1; Nordic, European and intercontinental origin). Treatments investigating impact of feed additives were excluded from the dataset, except for those based on terrestrial plant lipids which are commonly used in dairy cows' diet and are frequently represented in the database. The resulting database ($n = 99$, from 29 studies on dairy cows) is described in Table 1, where roughage and concentrate ratio and CH₄ production along with corresponding DMI are presented. The roughage was mainly comprised of silage from grass, maize and alfalfa, while barley, maize and soybean meal were the main ingredients of the concentrates. The CH₄ production was determined by the sulfur hexafluoride (SF₆) gas tracer technique in 14 studies, by respiration chambers in 13 studies, by the hood calorimetry technique in one study and by the GreenFeed system in one study.

Table 1. Summary of database for the basic models.

Data-Base ^a	Stage ^b	N ^c	Roughage	Concentrate	Forage Proportion (% of DM)	DMI (kg/day) ^d	CH ₄ Collection Technique ^e	CH ₄ (MJ/day) ^f	References
D	L	4	Maize silage	Ground maize	50	20	1	20 (14–26)	[10]
D	NL	4	Grass hay or barley silage	Barley grain	95	11	1	12 (11–17)	[11]
D	L	3	Grass silage	Oats, barley, peas and rapeseed cake	69	16	1	17 (16–18)	[12]
D	L	2	Grass silage	Barley, wheat and maize	73	23	1	32 (28–36)	[13]
D	L	3	Grass silage	Barley, wheat and oats	77	20	1	26 (24–28)	[14]
D	L	6	Ryegrass, white and red clover	Pelleted barley	77	19	2	24 (23–26)	[15]
D	L	3	Grass and maize silage	Barley	67	17	2	19 (17–21)	[16]
D	L	3	Alfalfa hay and alfalfa silage	Barley, maize and peas	51	26	1	23 (22–25)	[17]
D	L	4	Grass silage	Barley	70	17	1	25 (21–30)	[18]
D	NL	4	Grass silage	Wheat starch (non-NDF concentrate)	83	8	1	11 (10–12)	[19]
D	L	6	Grass silage	Wheat starch (non-NDF concentrate)	69	15	1	18 (17–19)	[20]
D*	L	4	Grass silage	Oats, barley and rye	50	19	1	26 (25–28)	[21]
D*	L	2	Rye grass, white clover or mature diverse pasture	0	100	21	4	27 (26–28)	[22]
D*	L	1	Grass clover silage	0	100	12	2	17	[23]
D*	L	1	Maize, grass/clover silage	Barley, sugar beet pulp and rapeseed cake	50	19	2	18 (16–20)	[24]
D*	L	2	Hay, maize silage and grass pellets	Wheat, maize, barley, rapeseed cake	80	21	2	27 (26–28)	[25,26]
D*	L	2	Maize and grass/clover silage	Whole cracked rapeseed	55	21	2	25 (23–27)	[27]
D*	L	6	Maize, grass silage and hay	Oat, soybean, wheat and apple pulp	50	17	2	22 (18–25)	[3]
D*	L	3	Ryegrass	0	100	15	2	17 (16–19)	[28]
E	L	4	Grass and maize silage	Rapeseed meal, rapeseed cake, cracked rapeseed	51	18	1	20 (17–23)	[29]
E	L	6	Grass silage and maize silage	Rapeseed meal, whole crushed rapeseed	64	17	1	20 (18–22)	[30]

Table 1. Cont.

Data-Base ^a	Stage ^b	N ^c	Roughage	Concentrate	Forage Proportion (% of DM)	DMI (kg/day) ^d	CH ₄ Collection Technique ^e	CH ₄ (MJ/day) ^f	References
E	L	4	Alfalfa hay and ryegrass silage	Cracked wheat grain	63	20	2	26 (25–28)	[31]
E	L	2	Maize and grass silage	Soybean meal and rolled barley	80	17	1	18 (14–22)	[32]
E	L	2	Maize silage and alfalfa haylage	Cracked wheat grain	67	16	1	23 (21–25)	[33]
E	L	4	Barley silage	Steam rolled barley and pelleted supplement	45	18	2	15 (13–16)	[34]
E	L	2	Haylage, maize silage and high moisture maize	Maize gluten and soybean meal	59	15	3	19 (15–23)	[35]
E	L	4	Hay, grass and maize silage	Barley and wheat bran Rapeseed meal,	75	17	2	22 (18–24)	[36]
E	L	4	Maize and grass silage	sunflower meal, ground wheat and maize gluten feed	56	20	2	23 (22–23)	[37]
E	L	4	Alfalfa silage	High moisture maize and dry maize	88	24	2	25 (24–26)	[38]

^a D, experiments used for model development; * indicates newly added studies; E, experiments used for model evaluation; ^b Physiological stage defined as either lactating (L) or non-lactating (NL); ^c Number of treatment means in study; ^d Mean value of dry matter intake (DMI) for experiment; ^e 1, tracer gas technique; 2, chamber; 3, head hood; 4, GreenFeed system; ^f Mean (min–max) value for experiment; the following factors were used in converting CH₄ in L/day to g/day and g/day to MJ/day: 1 L CH₄ = 0.716 g; 1 g CH₄ = 0.05565 MJ.

2.2. Development of Basic Models

CH₄ production was predicted by fitting mixed models to the lmer [39] procedure of R statistical language (R Core Team 2016; version 4.0.2) (Equation (1)):

$$Y = \beta_0 + \beta_1 X_1 + \beta_2 X_2 + \beta_n X_n + R_j + \varepsilon \quad (1)$$

where Y denotes the response variable of CH₄ production, β_0 denotes the fixed effect of intercept; X_1 to X_n denote the fixed effects of predictor variables and β_1 to β_n are the corresponding slopes; R_j denotes the random study effects of the experiment; ε denotes the within-experiment error. To account for differing accuracy in observed means, models were fitted using the WEIGHT statement in R (R Foundation for Statistical Computing, Vienna, Austria), where the data were weighted according to the number of observations [40]. The effect of the categorical factor CH₄ measurement techniques (tracer gas, chamber, headhood, GF) was included in the model as a fixed effect prior to final model development and found to be not significant ($p > 0.1$), and thus was not incorporated in the final models fitted. The presence of multicollinearity of fitted models was examined based on the variance inflation factor (VIF). A VIF in excess of 5 was considered an indicator of multicollinearity [41]. Multicollinearity was not detected. All parameters included in the developed models presented were significant at $p < 0.05$.

2.3. Basic Model Evaluation

In total, ten models were evaluated, including three models developed in the present study and seven extant models with similar input variables (DMI and dietary nutrient contents). The models were compared through assessing their abilities of predicting CH₄ production, using mean squared prediction error (MSPE) and concordance correlation coefficient (CCC). The MSPE was calculated according to Bibby and Toutenburg [42] as shown in Equation (2):

$$MSPE = \frac{\sum_{i=1}^n (Y_i - \hat{Y}_i)^2}{n} \quad (2)$$

where Y_i denotes the observed value of the response variable for the i th observation, \hat{Y}_i denotes the predicted value of the response variable for the i th observation, n denotes the number of observations. The root mean square prediction error (RMSPE) was used to assess overall model prediction accuracy because its output was in the same unit as the observations. In the present study, RMSPE was reported as a proportion of observed CH₄ production means in order to compare the predictive capability of models with different predicted means. A smaller RMSPE implies a better model performance. The MSPE was decomposed into error in central tendency (ECT), error due to disturbance (ED) or random error and error due to regression (ER).

The ECT, ED and ER fractions of MSPE were calculated as follows:

$$ECT = (\bar{P} - \bar{O})^2 \quad (3)$$

$$ED = (1 - R^2) \times S_o^2 \quad (4)$$

$$ER = (S_p - R \times S_o)^2 \quad (5)$$

where \bar{P} and \bar{O} are the predicted and observed means, S_p is the predicted standard deviation, S_o is the observed standard deviation and R is the Pearson correlation coefficient.

According to Lawrence and Lin [43], CCC is the product of a bias correction factor as the measurement of accuracy (C_b) and the precision measurement of Pearson correlation coefficient (r). The CCC was calculated as shown in Equation (6):

$$CCC = r \times C_b \quad (6)$$

where:

$$C_b = [(v + 1)/(v + \mu^2)/2]^{-1}$$

$$v = S_o/S_p$$

$$\mu = (\bar{P} - \bar{O}) / (S_o S_p)^{1/2}$$

where \bar{P} , \bar{O} , S_o and S_p were defined above, and v indicates a measure of scale shift, and μ indicates a measure of location shift. The CCC evaluates the degree of deviation of the best-fit line from the identity line ($y = x$), and thus, the CCC of a model that is closer to 1, is an indication of better model performance.

2.4. Update of Operational Models

The operational equation from Storlien and Harstad [44] presently used for predicting Y_m was based on calculations in NorFor (Table 2), using intervals of 500 kg from 5000 to 12,000 kg of ECM. The Norfor database with CH_4 production (not shown) predicted by the basic models, GEI and Y_m (not shown; calculated based on CH_4 production and GEI) was used in the present study for the update of operational models. The standardized lactation curves in NorFor were employed to predict animal requirement for ECM production through the lactation cycle. Daily DMI was calculated for every second lactation week for each 500 kg interval of the 305-day lactation. Feed energy (GE, metabolizable energy (ME) and net energy (NE)), animal energy requirements and energy supplementation were calculated based on the Dutch net energy lactation (NEL) system [45] as modified by NorFor [8].

Table 2. Mean (min-max) value of concentrate share, dry matter intake (DMI) and gross energy intake (GEI) throughout a 305-day lactation with various combinations of silages and concentrates at different levels of energy corrected milk (ECM) production ^a in the NorFor-database used for the operational models.

Yield (ECM, kg)	Silage ^b	Concentrate ^c	Concentrate Share, % DM ^d	DMI, kg/d	GEI, MJ/day
5000	1	I	11 (0–37)	15 (12–17)	279 (232–312)
	2	II	20 (0–53)	15 (12–17)	282 (228–327)
	3	II	25 (0–50)	16 (12–18)	292 (233–340)
5500	1	III	13 (0–40)	15 (13–17)	289 (242–323)
	2	III	16 (0–38)	16 (13–17)	292 (245–323)
	3	II	29 (10–51)	16 (12–19)	305 (232–355)
6000	1	III	14 (0–40)	16 (14–18)	300 (255–331)
	2	I	23 (3–47)	16 (14–19)	307 (253–352)
	3	II	32 (9–52)	17 (14–20)	319 (252–368)
6500	1	III	16 (0–43)	17 (14–18)	310 (261–342)
	2	III	22 (4–47)	17 (14–19)	316 (268–350)
	3	III	35 (11–52)	18 (14–20)	333 (267–383)
7000	1	II	21 (1–53)	17 (15–19)	324 (276–359)
	2	III	23 (7–45)	17 (15–19)	322 (276–354)
	3	II	39 (16–55)	19 (15–21)	347 (279–398)
7500	1	III	20 (4–47)	18 (15–19)	330 (284–362)
	2	I	32 (15–53)	18 (15–21)	345 (278–394)
	3	II	42 (21–57)	19 (16–22)	361 (292–412)
8000	1	III	22 (7–49)	18 (16–20)	340 (294–371)
	2	I	35 (17–54)	19 (16–22)	359 (291–407)
	3	II	45 (26–59)	20 (16–23)	376 (307–427)
8500	1	III	24 (10–50)	19 (16–20)	350 (303–383)
	2	I	37 (18–55)	20 (16–22)	372 (308–422)
	3	II	47 (30–61)	21 (17–24)	390 (320–442)

Table 2. Cont.

Yield (ECM, kg)	Silage ^b	Concentrate ^c	Concentrate Share, % DM ^d	DMI, kg/d	GEI, MJ/day
9000	1	III	26 (12–52)	19 (17–21)	360 (313–393)
	2	I	40 (21–57)	21 (17–23)	386 (319–436)
	3	II	50 (34–63)	22 (18–24)	405 (334–457)
9500	1	I	38 (23–59)	21 (17–23)	387 (315–437)
	2	I	43 (25–59)	21 (18–24)	400 (332–451)
	3	I	49 (35–61)	22 (18–25)	413 (346–464)
10,000	1	I	39 (23–60)	21 (18–24)	401 (332–452)
	2	I	45 (29–60)	22 (18–25)	414 (346–466)
	3	I	52 (38–62)	23 (19–25)	427 (358–477)
10,500	1	I	41 (23–62)	22 (19–25)	415 (348–467)
	2	I	48 (32–61)	23 (19–25)	429 (359–480)
	3	I	54 (41–64)	23 (20–26)	441 (370–491)
11,000	1	I	43 (25–63)	23 (19–26)	429 (358–480)
	2	I	50 (35–62)	24 (20–26)	443 (372–495)
	3	I	57 (43–67)	24 (20–27)	454 (381–504)
11,500	1	I	46 (29–64)	24 (20–26)	443 (373–496)
	2	I	52 (38–63)	24 (21–27)	457 (388–510)
	3	I	59 (46–70)	25 (21–27)	468 (393–518)
12,000	1	I	48 (32–65)	24 (21–27)	458 (387–511)
	2	I	54 (41–65)	25 (21–28)	472 (401–525)
	3	I	59 (48–68)	26 (21–28)	484 (404–537)

^a The standardized lactation curves in the NorFor-database were employed to predict animal requirement for ECM production through the lactation cycle; ^b 1, 2 and 3 refer to code for silages in Table 3; ^c I, II and III refer to code for concentrates in Table 3; ^d DM: dry matter. Silages 1, 2 and 3 represent a normal range in forage qualities found in the Norwegian cattle production; the combinations of silage and concentrate were determined on the basis of minimum cost when the energy requirements of the animal are met.

The data predicts standard feed rations during a 305-day lactation at different lactation yield, using three different forage qualities (Table 3), 5.7, 6.1 and 7.0 MJ NEL per kg DM, representing low, medium and very high energy content, respectively. Three complimentary concentrate mixtures, which are representative of what is used in practical diet formulation in Norway, were used in the diet formulation to meet the animal energy requirement (Table 3).

Table 3. Chemical composition (per kg of dry matter) of silages and concentrates in the NorFor^a-database used for the operational models.

Feed Type	Code	Nutritional Value	DM ^b (g/kg)	Ash (g)	Crude Protein (g)	Crude Fat (g)	NDF ^c (g)	Total Acids (g)	Sugar (g)	Starch (g)	Net Energy for Lactation (MJ)
Silage	1	Very high	332	77	167	39	436	62	92	n.d.	7.0
	2	Medium	325	70	157	35	511	63	53	n.d.	6.1
	3	Low	320	68	150	34	538	64	43	n.d.	5.7
Concentrate ^d	I	High	879	83	200	59	182	n.d.	n.d.	301	8.0
	II	Medium	873	76	194	52	208	n.d.	n.d.	307	7.7
	III	Low	873	76	182	46	202	n.d.	n.d.	390	7.5

^a NorFor: Nordic Feed Evaluation System [8]; ^b DM: Dry matter; ^c NDF: Neutral detergent fiber; ^d Concentrates with high (I), medium (II) and low (III) net energy content were FORMEL Energi Premium 80, FORMEL Elite 80 and FORMEL Favor 80, respectively (Felleskjøpet Agri, Lillestrøm, Norway); n.d.: not determined.

To observe the effects of different basic models on the output of operational models, the basic model that performed the best in predicting CH₄ production, and models from Storlien et al. [7] and Nielsen et al. [6] were selected to predict CH₄ production, respectively, and thus to calculate Y_m in the NorFor-database. Three operational models were therefore developed, in which the response variable was Y_m, and the input variables were ECM

and concentrate share in the diet. Moreover, GEI was also predicted with the same input variables. The Y_m and GEI were estimated by fitting a mixed effect model using the lmer [40] procedure of R statistical language (R Core Team 2016; version 4.0.2). The model employed is shown in Equation (7):

$$Y = b_0 + b_1X_1 + b_2X_2 + b_nX_n + S_j + \epsilon \quad (7)$$

where Y denotes the response variable of Y_m or GEI, b_0 denotes the fixed effect of intercept; X_1 to X_n denote the fixed effects of predictor variables and b_1 to b_n are the corresponding slopes; S_j denotes the repeated effect of days after lactation at each ECM production level; ϵ denotes the error within a lactation cycle. The presence of multicollinearity of fitted models was examined based on the VIF. A VIF in excess of 5 was considered an indicator of multicollinearity [41]. Multicollinearity was not detected. The following equation was used to calculate the CH_4 emission factor (EF) for 365 days, which can be used for estimating national CH_4 emissions when the number of animals is known:

$$\text{EF} = (\text{GEI} \cdot Y_m \cdot 365 \text{ days/yr}) / 55.65 \text{ MJ/kg CH}_4 \quad (8)$$

where EF denotes emission factor (kg CH_4 /head/year); GEI denotes gross energy intake (MJ/head/day); Y_m denotes CH_4 conversion rate, which is the fraction of gross energy in feed converted to CH_4 .

3. Results

3.1. Development and Evaluation of Basic Models

Models 1, 2 and 3, which were developed in the present study, and other extant models, are presented in Table 4 with results of model evaluations. The models were arranged in descending order of CCC. Overall, the developed models and models from Storlien et al. [7] and Nielsen et al. [6] performed better than other extant models with respect to prediction accuracy (*RMSPE* and *CCC*), except that the lowest *RMSPE* was found in one of the models from Niu et al. [9] yet with low *CCC*. The overall performance of the extant models using only DMI as input variable did not perform as good as models where dietary FAs and/or NDF were included as input variables in addition to DMI. Model 1 slightly outperformed the model from Storlien et al. [7], judged by *RMSPE* (15.0 versus 15.3), owing to smaller *ER*. When NDF together with DMI and FAs was included as input variables in the models, evaluation through *CCC* and *RMSPE* indicated that model performances were improved (Model 2 and 3, as well as the Nielsen et al. [6] model). Model 2 and 3 performed even better, indicated by lower *RMSPE* and higher *CCC*, compared to the Nielsen et al. [6] model. It was assumed that cows are not emitting nor inhaling CH_4 if they are not eating, hence the intercept was forced to zero in Model 2 to have Model 3 developed. The performance was somewhat compromised for Model 3 as compared to Model 2 mainly due to increased *ED* (Table 4).

Table 4. Evaluation of developed and extant basic models ordered by decreasing CCC.

Model	<i>n</i>	Prediction Equation	RMSPE, %	ECT, %	ED, %	ER, %	CCC	<i>r</i>	<i>C_b</i>
Model 2	36	$CH_4 = -3.01 + 1.19 \times DMI - 0.103 \times FAs + 0.017 \times NDF$	13.8	0.2	86.1	13.7	0.703	0.70	1.00
Model 3	36	$CH_4 = 1.13 \times DMI - 0.114 \times FAs + 0.012 \times NDF$	13.9	0.1	87.3	12.6	0.694	0.69	1.00
[6]	36	$CH_4 = 1.23 \times DMI - 0.145 \times FAs + 0.012 \times NDF$	15.3	3.1	73.1	23.8	0.677	0.69	0.99
Model 1	36	$CH_4 = 4.92 + 1.13 \times DMI - 0.118 \times FAs$	15.0	0.9	82.8	16.3	0.650	0.65	1.00
[7]	36	$CH_4 = 6.80 + 1.09 \times DMI - 0.15 \times FAs$	15.3	0.6	79.3	20.1	0.649	0.65	1.00
[9]	36	$CH_4 = 26.0 + 15.3 \times DMI + 3.42 \times NDF/10 \times 0.05565$	13.0	0.0	97.6	2.40	0.611	0.70	0.87
[46]	36	$CH_4 = (38.0 + 19.22 \times DMI) \times 0.05565$	15.6	5.2	89.0	5.80	0.547	0.58	0.95
[9]	36	$CH_4 = [160 + 14.2 \times DMI - 13.5 \times EE/10] \times 0.05565$	15.6	14.8	84.0	1.20	0.528	0.60	0.87
[9]	36	$CH_4 = (107 + 14.5 \times DMI) \times 0.05565$	14.8	0.7	99.2	0.00	0.504	0.58	0.87
[47]	36	$CH_4 = (20 + 35.8 \times DMI - 0.5 \times DMI^2) \times 0.716 \times 0.05565$	15.4	8.2	90.9	0.90	0.434	0.57	0.76

n, number of treatment means; CH₄, methane (MJ/day); DMI, dry matter intake (kg/day); EE, ether extract content (g/kg dry matter; FAs, fatty acid content (g/kg DM); NDF, neutral detergent fiber content (g/kg DM) if not indicated otherwise; RMSPE, root mean squared prediction error expressed as a percentage of the observed mean and in MJ; ECT, error due to bias, as a percentage of total MSPE; ER, error due to regression, as a percentage of total MSPE; ED, error due to the disturbance, as a percentage of total MSPE; CCC, concordance correlation coefficient; *r*, Pearson correlation coefficient; *C_b*, bias correction factor.

Plots of observed versus predicted values of enteric CH₄ production and the residuals (observed minus predicted) for Model 3 and models from Storlien et al. [7] and Nielsen et al. [6] are presented in Figure 1. These three models were selected to calculate CH₄ production in the NorFor-database, respectively.

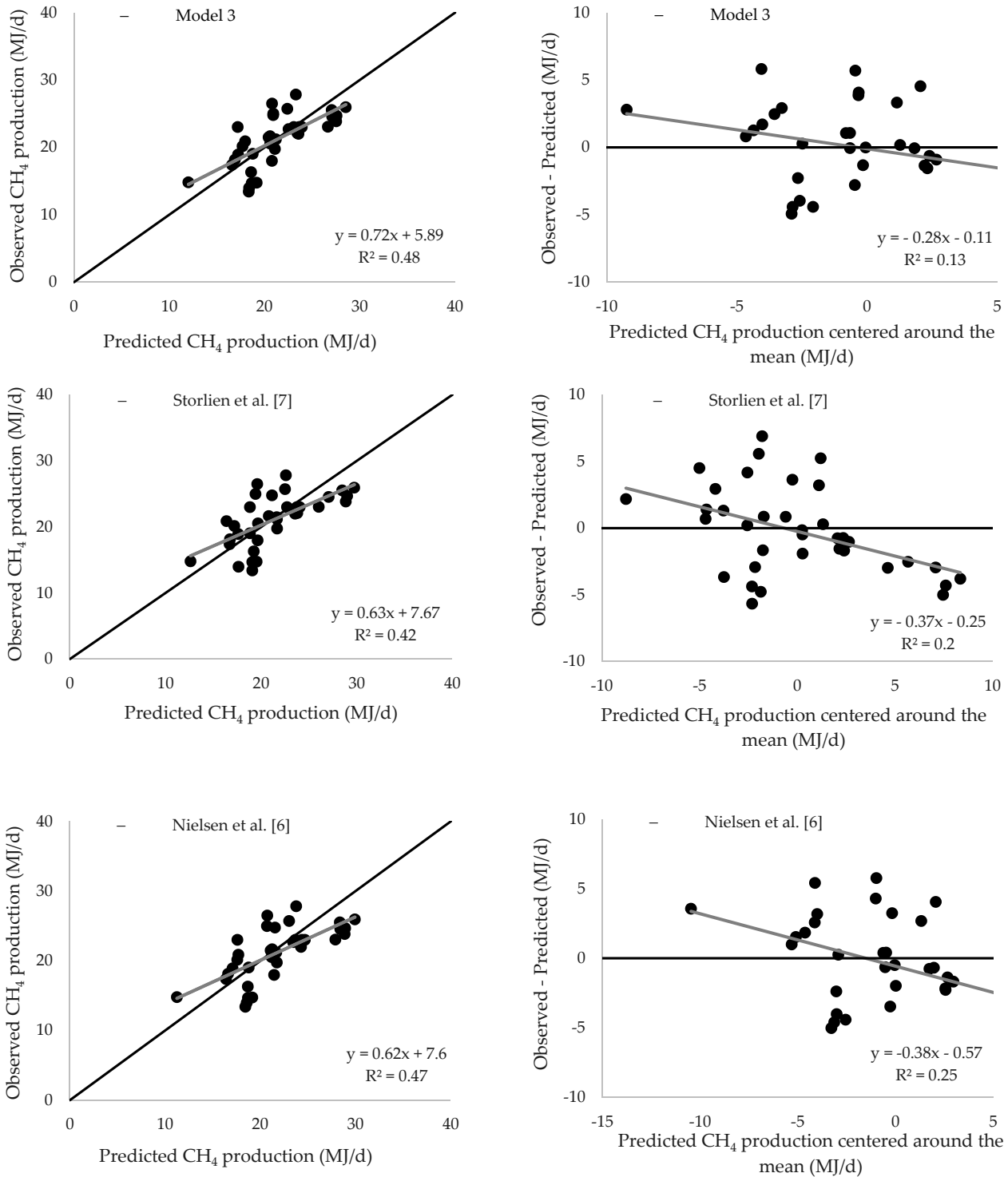


Figure 1. Observed versus predicted values of enteric CH₄ production and the residuals (observed minus predicted) for basic models used in Norway and the Model 3 developed in the present study. The graphs to the left show that the models overestimate CH₄ emissions at the lower range and underestimate emissions at the upper range. The graphs to the right show the presence of a linear bias (slope) and the presence of a mean bias (intercept).

3.2. Update of Operational Models

The operational models for the prediction of Y_m and GEI are presented in Table 5. There was a significant positive relationship between GEI and both ECM and concentrate share. When estimating Y_m , both predictor variables were negatively correlated to the response variable.

Table 5. Operational models: CH₄ emission factors (kg/year per cow), Y_m and gross energy intake (GEI), estimated using selected basic models at production levels of 6000, 8000 and 10,000 kg energy corrected milk (ECM) assuming 38.0, 43.5 and 50.0% concentrate share in the rations, respectively.

Model ^a	CH ₄ , kg/Year Per Cow ^b	Y_m ^c , %	GEI ^d , MJ/Cow and Day
GEI = 159 + 0.02 × ECM + 1.39 × conc.share			
6000 kg ECM and 38.0% concentrate share			
$Y_{m(S)} = 7.11 - 7 \times 10^{-5} \times \text{ECM} - 4.1 \times 10^{-3} \times \text{conc.share}$	127.7	6.53	298
$Y_{m(M)} = 7.65 - 1.1 \times 10^{-4} \times \text{ECM} - 5.4 \times 10^{-3} \times \text{conc.share}$	130.2	6.66	298
$Y_{m(N)} = 7.71 - 1 \times 10^{-4} \times \text{ECM} - 4.4 \times 10^{-3} \times \text{conc.share}$	131.5	6.72	298
8000 kg ECM and 43.5% concentrate share			
$Y_{m(S)} = 7.11 - 7 \times 10^{-5} \times \text{ECM} - 4.1 \times 10^{-3} \times \text{conc.share}$	146.5	6.40	349
$Y_{m(M)} = 7.65 - 1.1 \times 10^{-4} \times \text{ECM} - 5.4 \times 10^{-3} \times \text{conc.share}$	147.8	6.45	349
$Y_{m(N)} = 7.71 - 1 \times 10^{-4} \times \text{ECM} - 4.4 \times 10^{-3} \times \text{conc.share}$	150.6	6.57	349
10,000 kg ECM and 50.0% concentrate share			
$Y_{m(S)} = 7.11 - 7 \times 10^{-5} \times \text{ECM} - 4.1 \times 10^{-3} \times \text{conc.share}$	164.5	6.25	401
$Y_{m(M)} = 7.65 - 1.1 \times 10^{-4} \times \text{ECM} - 5.4 \times 10^{-3} \times \text{conc.share}$	163.7	6.22	401
$Y_{m(N)} = 7.71 - 1 \times 10^{-4} \times \text{ECM} - 4.4 \times 10^{-3} \times \text{conc.share}$	168.2	6.39	401

^a $Y_{m(S)}$, $Y_{m(M)}$ and $Y_{m(N)}$ denotes Y_m calculated based on GEI (Norfor-database) and CH₄ production which was predicted using the model from Storlien et al. [7], Model 3 and the model from Nielsen et al. [6], respectively; ^b Including 60 day of dry period through inclusion of dry cows in the model for predicting daily CH₄ production (MJ); ^c Y_m , methane conversion factor (% of GEI); ^d GEI: gross energy intake.

Table 5 shows the annual production of CH₄ assuming an annual milk yield of 6000, 8000 and 10,000 kg ECM and an averaged concentrate share of 38.0, 43.5 and 50.0%, respectively. These are typical concentrate shares in Norway where concentrate is used on all dairy farms. When milk yield and concentrate share were increased, Y_m was predicted to decrease in all models, whereas GEI and the CH₄ emission factor were predicted and calculated to increase, respectively. At a production level of 6000 kg ECM and a 38% concentrate share, when the prediction of Y_m was obtained through the model from Storlien et al. [7], the prediction of $Y_{m(S)}$ (see footnote to Table 5) and the CH₄ emission factor (127.7 kg/year per cow) were the lowest. On the contrary, using the model from Nielsen et al. [6] to predict CH₄ production and Y_m under the same conditions with the NorFor-database led to the highest predicted values of both $Y_{m(N)}$ (see footnotes to Table 5) and the CH₄ emission factor. The same ranking for both Y_m and the CH₄ emission factor was found at a production level of 8000 kg ECM and a 43.5% concentrate share, while the differences among predictions of $Y_{m(S)}$, $Y_{m(M)}$ (see footnotes to Table 5) and $Y_{m(N)}$ were decreased. At a production level of 10,000 kg ECM and a 50% concentrate share, predictions of $Y_{m(M)}$ and correspondingly the CH₄ emission factor were the lowest, which were 6.22 and 163.7 kg/year per cow, respectively.

4. Discussion

The aims of the present study were to develop a basic model which can be used as a method for the accurate calculation of enteric CH₄ emissions from individual dairy cows, and to update the existing operational model for the prediction of Y_m and the CH₄ emission factor to be used in the national GHG inventory in Norway.

4.1. Relationship between Methane Production and Dietary Factors in the Basic Models

In the present study, DMI and dietary concentrations of FAs and NDF were used and confirmed as key predictor variables for CH₄ production in dairy cows. DMI was the most important variable for the prediction of enteric CH₄ production in all models evaluated. The significant positive relationship is consistent with the knowledge that CH₄ production increases with feed intake due to the greater availability of substrate for microbial fermentation [8,48,49]. A linear relationship between DMI and CH₄ production has been observed in many studies [6,7,46]. However, an increased intake potentially increases passage rate of feed through the rumen, resulting in a decline in rumen fermentation and CH₄ production per unit of feed [50]. Subsequently, the percentage of gross energy lost as CH₄ declines [9], but at the same time digestibility may decline resulting in an unchanged methane emission intensity per unit of milk or meat produced. Nevertheless, the first assumption implies that in theory a model of CH₄ production based on DMI, GEI or MEI, should be nonlinear [8]. The only nonlinear model [47] that was evaluated in the present study did not perform as robust as others, which may be due to that only feed intake was accounted for in their model. This could be justified by Bell et al. [51], where the residual variation (difference between observed and predicted values) in CH₄ emission was notably reduced after incorporating the significant fixed effects of dietary characteristics on CH₄ yield, in addition to the effect of feeding level.

Fat content was the second most important variable for the prediction of enteric CH₄ production in all models evaluated. In the present study, the accuracy of prediction was better with the inclusion of dietary fat content in the equation compared to extant models where only DMI was used, and there was a significant negative relationship between fat and CH₄ production. This was facilitated by not excluding experiments where fat had been supplemented. Indeed, CH₄ production decreases through fat supplementation in the diet, as reviewed and studied by several groups [11,34,51]. The mode of action of fat on CH₄ mitigation has been extensively studied [52]. The effect is based on the following components. (1) Biohydrogenation of unsaturated fatty acids utilizes H₂ available for CH₄ production. However, the complete biohydrogenation of one mol of linoleic acid can reduce CH₄ production only by one mol and thus this is not quantitatively important [47]. (2) As fat is not fermentable, part of the reduced CH₄ production with increased dietary fat concentration can be accredited to decreased supply of fermentable substrate for the microorganisms, also reducing hydrogen production [53]. (3) The most important component is a direct toxicity of fatty acids, especially that of lauric and myristic acid and polyunsaturated fatty acids, exhibiting against the archaeal methanogens [54]. (4) Finally, dietary fat concentration directly influences rumen fermentation by favoring propionate production at a cost of acetate or butyrate, or both, because protozoa are inhibited as well which results in declines in fiber digestion and hydrogen supply [55].

The accuracy of prediction was further improved when dietary NDF content was included in the equations along with DMI and fat, and there was a significant positive relationship between NDF and CH₄ production as expected from earlier studies [6,56]. Studies focusing on the effect of different types of carbohydrates, indicate that high concentrations of starch and sugar (non-fibrous carbohydrates) increase the production of propionate but decrease that of acetate and butyrate, and the opposite is true for NDF (fibrous carbohydrates) [53,56]. The CH₄ production is thus related to the VFA profile in such a way that higher NDF increases CH₄ production by shifting short chain fatty acid proportion towards acetate which is associated with a higher hydrogen release [57]. The NDF content was only the third most important variable for the prediction of enteric CH₄ production in all models evaluated, i.e., the influence of NDF content was less pronounced than that of fat contents.

Model 3 was developed from Model 2 by applying biologically sensible constraints, e.g., zero CH₄ at zero intake [8]. In the current study, Model 3 was selected based on model performance as the updated model over models from Nielsen et al. [6] and Storlien et al. [7]. Different from the Storlien et al. [7] equation, Model 3 allows for considering

effects of NDF concentration in the feed in addition to fat concentration. The concentration of NDF will vary with forage proportion and quality in the diet. A positive coefficient for NDF reflected reduced CH₄ production by earlier harvesting of grass for silage as NDF concentration in grass increases with harvesting time. Model 3 has the same input variables as the Nielsen et al. [6] equation but yields slightly lower estimates of the comparatively high CH₄ emission factor in Norway (Table 5).

4.2. Update of Operational Models

The NorFor-database applied in the present approach is the same as used by Storlien and Harstad [44], and the calculation of GEI remained unchanged. No major changes in milk yield and quality of silage and concentrate have taken place since 2015 (pers. com. TINE and Felleskjøpet Fôrutvikling), and therefore, it was considered unnecessary to recalculate the NorFor-data, except CH₄ production. However, since input data of predicted enteric CH₄ production was changed, equations for prediction of Y_m based on ECM and concentrate share also changed. Many studies have suggested using factors such as fiber digestion [58,59] and dietary lipid content [60], either as the single or multiple variables of a Y_m model. However, in the present study a country-specific approach was used for the prediction of Y_m using the same method as Storlien and Harstad [44]. This approach allows country-specific information to be included in the development of equations without access to data that are not readily available, such as fiber and lipid contents in the diet. In the Norwegian cow recording system (CRS) individual milk yield and concentrate supplementation is reported 11 times per cow per year, and data from 1.16 million individual cow observations are available [8]. The recorded information in the Norwegian CRS was not directly included for updating the operational models. Instead, the simulated Norfor-database (Table 2) included a variety of variables such as feed intake and composition, Y_m and GEI, in addition to milk yield and concentrate share. In order to develop representative Y_m for the about 200,000 Norwegian dairy cows this was essential for being able to take into account the effect of dietary composition and the experiments using grass-based diets, which were considered when updating CH₄ production in the NorFor-database. From Table 5 the predicted Y_m, depending on the level of production, ranged from 6.22 to 6.72%, which is within the range of the IPCC default Y_m of 6.5% ± 1% [61]. This default value is recommended by IPCC [61] for all types of cattle and buffalo, except feedlot cattle fed at least 90% concentrate. However, the lowest predicted value 6.22% was yet higher than that given by Hellwing et al. [62] for Danish dairy cows, which was 6.02% and 5.98% of GE intake for Holstein and Jersey cows, respectively. Accordingly, Lesschen et al. [63] concluded that within the EU countries, the GHG emission per kilogram milk produced was lowest in Denmark. In the Netherlands, a Tier 3 approach which addresses effects of nutritional details on enteric CH₄ emission is used for the national inventory, with a predicted CH₄ emission factor in a smaller range of 110.5 to 129.4 kg/cow/year and a lower predicted Y_m of 5.88% to 6.07% of GE intake [64] at unspecified production level. In France, a new equation was developed to predict enteric CH₄ that complies with IPCC rules for a Tier 3 method and is based on digestible organic matter intake (DOMI). The representative dairy cow of 650 kg BW and 6300 kg annual milk yield was estimated to produce only 119.3 kg CH₄/year using a default Y_m value of 6.50% [65], while the operational model of the present study yields as much as 130 kg CH₄ per year at a production level of 6000 kg ECM/year. The discrepancies across countries can possibly be explained by differences in diet composition, as there is a higher dietary proportion of forage in Norway, and milk yield is moderate compared to other European countries and USA. With increasing milk yield and concentrate share, Y_m decreases, whereas the CH₄ emission factor increases. This is due to the fact that more energy is allocated to milk production, as the CH₄ emission in kg per kg ECM decreased. These results are in accordance with those reported by Kirchgessner [66] and Volden and Nes [8]. Accordingly, CH₄ emission decreases by 2.8 g/kg milk and 41.4% of total

CH₄/milk per day when milk production is increased from 4000 to 6000 kg and from 5000 to 9000 kg, respectively.

The value of operational models is dependent on correct and annually updated reporting of average annual milk yield and concentrate share of dry matter intake. In addition, an updated basic model could help refining the estimates of CH₄ production, which could ultimately improve the estimate of Y_m . As discussed above, it is possible by using the above information to develop a robust model for use in Norway for the calculation of enteric CH₄ emission from dairy cows. Further, the recommended equation is well suited for improving the CH₄ emissions estimates of the farm level net GHG model HolosNor [67]. The HolosNor is used as an advisory tool [68], and the implementation of Model 3 developed in the current work will be helpful for quantifying and advising mitigation strategies at farm level. In the current models developed, the effects of dietary changes were considered only indirectly through calculation of Y_m using basic models. Therefore, a further improvement in the prediction accuracy might be expected for a tier 3 model that includes also a dynamic and mechanistic model of fermentation biochemistry to calculate enteric CH₄ emission inventories [64,65].

5. Conclusions

Three basic models were developed in this study. Among them, Model 3 with input variables of DMI, dietary concentrations of FAs and NDF, turned out to predict CH₄ production more accurately than the extant models from Nielsen et al. [6] and Storlien et al. [7]. Using a basic model database containing recently published data improved CH₄ production estimates in the operational model. Hence, this basic (Model 3) and updated operational equation for calculation of enteric CH₄ emission from individual dairy cows in Norway is now used by the Norwegian Environment Agency (Miljødirektoratet). This is essential to improve accuracy of carbon footprint assessment of dairy cattle production systems and to help quantify and communicate effective mitigation strategies.

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